

The Product
Carbon
Footprint
Guideline for
the Chemical
Industry

Specifications for suppliers' product carbon footprint calculation

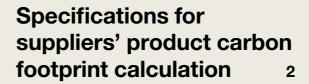
# The PCF Guideline document contains 5 chapters.

The first version of the PCF Guideline focuses exclusively on chapter 5 of the Guideline, prescribing the specifications for supplier product carbon footprint (PCF) calculations within the chemical industry. The three additional chapters – "About the guidance" (chapter 2), "Reporting principles" (chapter 3) and "Guidance on scope 3.1 calculation on corporate level" (chapter 4) – will be published in November 2022.

TfS published chapter 5 separately in September 2022 because it recognizes the urgent need for a harmonized PCF calculation approach across the chemical industry. This is because a major share of the industry's greenhouse gas (GHG) emissions arises from the upstream value chain (scope 3). Increasing data transparency and accuracy on the product-level is absolutely critical to drive emission reductions along the value chain and is a strategic cornerstone of many corporate climate mitigation strategies. The TfS PCF guideline is unique in that it draws on the wealth of expertise and knowledge within the TfS member network to set a standard for the chemical industry, while remaining fully compliant with existing standards including ISO and the Greenhouse Gas Protocol. The PCF guideline will create benefits for TfS members, their suppliers, as well as other industries initiatives as a drop-in solution for the chemical sector.

In the subsequent release of this document, TfS members and their suppliers will be able to holistically approach the integration of PCFs of chemical products within their corporate GHG inventories, with a focus on Scope 3 Category 1 (purchased goods and services) emissions. The comprehensive guideline will instruct companies on how to calculate their own corporate inventories on the basis of supplier-specific data, while at the same time providing guidance on how to calculate the PCFs of their own chemical products, with the aim to create transparency and decarbonize the entire value chain.





5.1	Goal and Scope	4
5.1.1	General	4
5.1.2	System boundaries	5
5.1.3	Declared unit (DU) of PCF	6
5.2	Calculation rules	7
5.2.1	Steps of PCF calculation	7
5.2.2	Temporal Scope	7
5.2.3	Criteria to exclude certain activities (Cut-off)	8
5.2.4	Standards used	8
5.2.5	Data types and sources	9
5.2.6	Emission factor requirements and sources	10
5.2.7	Life Cycle Impact Assessment (LCIA)	11
5.2.8	Activity data requirements	13
5.2.9	Multi-output processes	29
5.2.10	Additional rules and requirements	36
5.2.11	Data quality and share of primary data	45
5.3	Verification and Reporting	50
5.3.1	Verification of PCF calculations / Quality Assurance	50
5.3.2	Information to be reported with PCF	51









Product-level CO<sub>2</sub> transparency along the value chain is crucial to identify, track, and reduce greenhouse gas (GHG) emissions in cooperation with supply chain members.

This transparency is increasingly demanded by customers from all industrial sectors who are strongly and increasingly targeting the reduction of GHG emissions.

The sharing of Product Carbon Footprints (PCF) information between supply chain members enables companies to track their scope 3 GHG emissions and facilitate reduction efforts [GHG Protocol Scope 3 Standard (2011)].

The following requirements apply to the calculation of product-related cradle-to-gate GHG inventories and serve as a global standard/guideline for calculating PCFs in the chemical industry. Adhering to these requirements enables comparability in PCF calculations and hence a level playing field. To create greater transparency and enable comparability, information on the exact methods or standards applied shall be shared downstream as part of the elements for data exchange.

The guideline is applicable to all chemical products, independent of their final use.

PCFs are modelled according to comparative guidelines/ standards, providing consistency in how the results have been modelled. The PCF-result between two comparable materials may differ because of differences in technologies, data used from suppliers, geographical aspects, etc.

However, the basis for the modeling should be well described and related to guidelines such as this one to avoid differences that come from using different assessment approaches. The calculation of results should be linked to a meaningful and harmonized reporting that explains in which way the calculations were executed and on which basis the results were generated, specifically in cases of the application of a variety of different methods. Furthermore, the calculation basis, specifically in cases of the application of a variety of different approaches shall follow this guideline. The practitioner or the persons in charge of the creation of the PCF are responsible for the preparation, calculation, quality, and the reporting of the PCF to a third party.

The calculation is only auditable if the reporting is done by the supplier accurately. Therefore, an attributes list and specific requirements were added to this document to enable data exchanges via specific platforms and to ensure that the recipient gets clear, high quality and meaningful information.

The guideline was prepared by experts of the "Together for Sustainability (TfS)" organization together with testing companies and third-party organizations. It reflects the status quo of the main recognized standards in the world. It was specified by requirements, procedures, assessment approaches for chemicals. The guideline will be updated if significant changes or adaptations are needed due to changes of other generic standards, new aspects that have not been considered so far or new requirements from the market. It will be published after indicating the revision on the TfS webpage with the changes that have been made compared to the previous version. The outdated versions will be stored in an accessible archive of TfS.

TfS recognizes that it is often difficult to compare PCF data of similar products because of the different underlying methodological decisions made in the calculation, uncertainties of data used, different levels of quality of data, differences in regions, technologies etc. However, the application of this guideline aims to reduce the issues to compare PCF of chemicals. In the future, PCFs will be important information sources to support companies in their GHG reduction strategies.

PCF information from suppliers in accordance to a sectorspecific guideline will contribute to the transparency along supply chains. A good reporting addressing all relevant information e.g., scope, standards used, PCR applied, data sources used, allocation methods applied, etc. will allow a better understanding of PCF results for chemicals.

The purpose of the PCF study report is to describe the PCF study, including the PCF or the partial PCF, and to demonstrate that the provisions of this document have been met. The PCF results generated by the companies can be used in different ways. The first instance is a B2B exchange of the data with an internal review recommended. Furthermore, the companies can publish PCF results in different ways, where an external review is requested [ISO 14026:2017]. The results and conclusions of the PCF study shall be documented in the PCF study report without bias. The results, data, methods, assumptions, and the life cycle interpretation shall be transparent and presented in sufficient detail to allow the reader to comprehend the complexities and trade-offs inherent in the PCF study [ISO 14067: 2018].

This guideline focusses on all relevant GHGs as defined by the Intergovernmental Panel on Climate Change (IPCC). The relevant GHG emissions and their emission factors are described in detail in 5.2.6.

However, the general principles can be used and applied for chemicals as well, if other environmental impacts beyond GHGs (e.g., air quality, water use, biodiversity) need to be addressed. These questions are becoming a more and more common ask from customers of the chemical industry and a leverage of the same method across impacts can be possible. Further specifications are needed in this context and can be seen as a possible future task resulting in an extension of the guideline.

In Figure 5.1 an overview is given for easier navigation in the guideline document and to more easily find the most relevant chapters and skip others. Figure 5.1 should also give support for beginners in this topic to start relatively quick with the first calculations and follow-up with specific questions later if relevant.

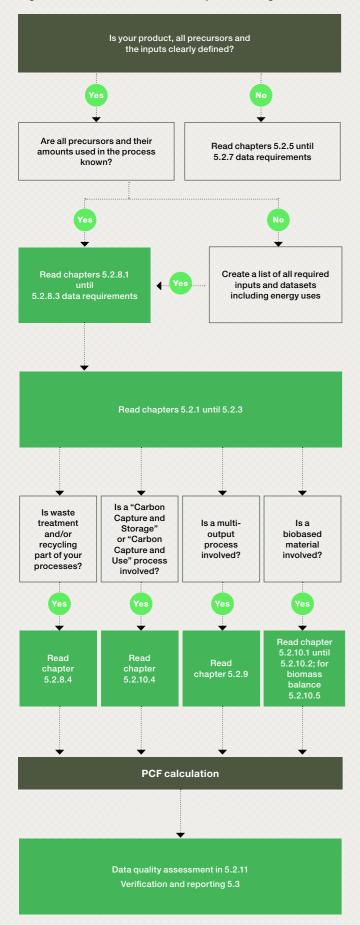
Currently, TÜV Rheinland Energy GmbH is providing the following services to TfS, which are expected to be completed in Q3/Q4 2022:

- Assess the guideline against all relevant standards applied (e.g. SBTi, WBCSD, GHG Protocol etc.).
- Check if reporting requirements for applicants are sufficiently defined in the guideline.
- Test the level of usability and giving hints for optimization.
- Loops of discussions and potential improvements during testing stage (WP 1-4 of TfS) and finalization stage (WP 1-5 of TfS).

It can been confirmed that the approaches used and the calculation methodology are reasonable, transparent and appropriate for the purpose of the guideline. The presented approach as well as the calculation examples are coherent, transparent and comprehensible.



Figure 5.1 Overview of the main chapters of the guideline



### 5.1 Goal and Scope

#### 5.1.1 General

The scope of this guideline covers the so-called "cradle-to-gate"- approach to calculate a PCF and refers to a "declared unit" (see 5.1.3).

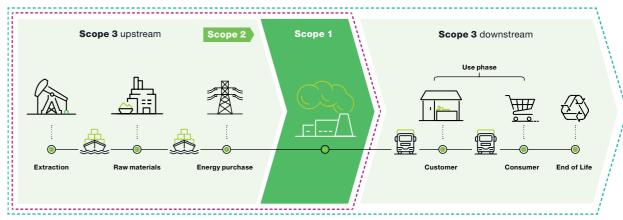
The guideline enables calculating the cradle-to-gate PCF based on standards and guidelines that were developed from different organizations.

General topics follow the standards mentioned in 5.2.4. It is stated, where the guideline defined specific rules for chemicals that are not reflected in detail in the current standards. The guideline is fully compliant with ISO and GHG Protocol. It is a challenge to be fully compliant with all other standards or guidelines that might be relevant. TÜV Rheinland checked and validated the compliance.

A Cradle-to-Gate PCF as used throughout this document, is the sum of GHG emissions and removals of one or more selected process(es) in a product system, expressed as  ${\rm CO}_2$  equivalents ( ${\rm CO}_2$ e) and based on the selected stages or processes within the life cycle. The selected stages in this guideline cover all activities within the defined system boundaries as defined in detail in Chapter 5.1.2.

It must be noted that a product assessment limited to only GHGs has the benefit of simplifying the analysis and producing results that can be clearly communicated to stakeholders. The limitation of a GHG-only inventory is that potential trade-offs or co-benefits between environmental impacts can be missed. Therefore, the results of a GHG-only inventory should not be used to communicate the overall environmental performance of a product [GHG Protocol Product Standard (2011)].

Figure 5.2 System boundary definition



Cradle-to-gate

Cradle-to-grave

### 5.1.2 System boundaries

The boundary of the guideline is a cradle-to-gate PCF, comprising all processes of extraction, manufacturing, and transportation, until the product leaves the factory gate. Downstream emissions from product use and end-of-life are in general excluded from a cradle-to-gate PCF (Figure 5.2).

The following activities **shall be included** in a cradle-to-gate PCF calculation: all product related direct (Scope 1) and indirect (Scope 2) GHG emissions of the production process, including fossil or biogenic removals, energy consumption (Scope 2: electricity, external heat and steam; Scope 1: fuel consumption like natural gas, biogas), utilities, manufacturing, inbound transportation, site-to-site transportation, treatment of process waste and wastewater treatment and all "Scope 3" related GHG emissions of raw material consumption including catalysts that are consumed in the reaction [BASF SE 2021]. Further information on included activities is provided in Table 5.1.

As the guidance is product-related, the following activities are **shall not be included** within the boundaries of a cradle-to-gate PCF: manufacturing of production equipment, buildings, infrastructure and other capital goods, business travel by personnel, travel to and from work by personnel, and research and development activities. [Pathfinder Framework (PACT powered by WBCSD)], Table 5.1. Please also see Chapter 5.2.3 on requirements to cut off activities.

The following activities might be included or excluded in the system boundary depending on cut-off criteria or customer requirements: Outbound transportation of the product is in general excluded (see Figure 5.2). If outbound transportation needs to be considered by customers' requests, it may be calculated and reported separately. Packaging of the product in question might be included or excluded. For many chemicals, the contribution of packaging may be negligible within the context of a PCF in terms of mass and environmental significance. This is for example the case for bulk chemicals which are delivered by a supplier to customer manufacturing sites. For other chemicals, such as specialty chemicals or construction chemicals, packaging

can play a more significant role in the PCF, namely for products sold in smaller units (e.g., in pails, cartridges, or wrapped rolls). In accordance with the cut-off criteria defined in section 5.2.3 of this guidance, packaging may be excluded or included in the PCF calculation, depending on its mass contribution and environmental significance. If packaging is included, it should be visible in the description of the declared unit (see 5.1.3).

The system boundary shall be the basis used to determine which unit processes are included within the PCF study. Where PCF Product category rules (PCR) are used, their requirements on the processes to be included supersede those indicated above (see 5.2.4). According to ISO 14067 [ISO 14067: 2018], a PCR is a "...set of specific rules, requirements and guidelines for carbon footprint of a product or partial carbon footprint of a product quantification and communication for one or more product categories". The criteria, e.g., cut-off criteria (5.2.3), used in establishing the system boundary shall be identified and documented internally in the PCF calculation report.

Decisions shall be made regarding which unit processes to include in the PCF study and to which level of detail these unit processes shall be analyzed. The exclusion of life cycle stages, processes, inputs, or outputs is only permitted if they do not significantly change the overall conclusions of the PCF calculation. In a "cradle to gate" approach, the use and disposal phases are not always of minor relevance but are not in the scope of the analysis and are excluded. In Chapter 5.1.3 the cut-off approach is described in detail. A variation of 10% of the PCF result by including or excluding life cycle stages is a variation that is generally accepted by practitioners due to inherent uncertainties, variabilities of factors or data sets used in a PCF calculation. Any decisions shall be clearly stated in the internal PCF calculation report and the reasons and implications for their exclusion shall be explained. The threshold for significance shall be stated and justified (see guidance in chapter 5.2.3).

The following table describes generically the activities that shall be included or excluded from the system boundaries as well as the ones that are optional.

5



Table 5.1 Activities to be included and excluded in the system boundaries and optional activities

Included	Excluded	Optional
Production related raw materials (including catalysts and ancillary materials that are consumed) <sup>1</sup>	Services such as engineering or infrastructure services, R&D activities	Packaging depending on the specific product and fulfilment of cutoff requirements
Utilities consumed	Business travel or employee commuting	Outbound transportation (if included in system boundary, it shall be stated separately)
Energy consumption	Production of investment goods	In-bound transportation if not relevant
Direct emissions from manufacturing and related on site utilities production/generation	Activities falling under the cut-off requirements (as provided in Chapter 5.2.3)	
Transportation of raw materials and site-to-site transportation		
Treatment or disposal of process wastes and wastewater treatment		

(1) Non-production-related procurement (often called indirect procurement) consists of purchased goods and services that are not integral to the company's products but are instead used to enable operations. Non-production-related procurement may include capital goods, such as furniture, office equipment, and computers. Source: GHG Protocol Corporate Value Chain Standard.

### 5.1.3 Declared unit (DU) of PCF

The declared unit (DU) describes the quantity of a product that is used as the reference unit in the quantification of the Cradle-to- Gate PCF. In case of chemical products, the declared unit is often defined as 1 kg of product.

This TfS guideline deals exclusively with the use of a declared unit as it only guides in calculating Cradle-to-Gate PCFs and thus does not include the full product life cycle.

The PCF, expressed in kg CO<sub>2</sub> equivalents per declared unit, reflects the cumulated climate change impact of air emissions of greenhouse gases (GHGs). Every supplier of the same product shall calculate its emissions using the same declared unit [BASF SE 2021].

Standard unit should be kg CO<sub>2</sub> equivalents per kg product preferably. For some specific products like gases (e.g., Hydrogen, LPG) the PCF might be expressed per unit norm cubic meter of product. Furthermore, some products are sold based on a volume unit (like liter), and in that case the PCF might be expressed per volume unit. In these cases, conversion factors (densities with associated conditions) shall be provided by the supplier for conversion to kg which is required in the attributes list in chapter 5.3. Any other unit of measurement like pieces or Euro shall not be used.

For processes, the PCF may be expressed as kg CO<sub>2</sub>e equivalents per ton of distilled product, per ton of treated wastewater or per ton of product in a crystallization process.

Some sectors may use pieces or other units in the declared unit. Regardless of what is used, a sufficient physical transfer shall be communicated to be able to convert these units into kg.

The results of a PCF linked to the declared unit should be reported as  $\log{\rm CO_2}$  equivalents per declared unit with one decimal. More decimals are not meaningful due to the variability of the figures. Results with a second decimal should be rounded: In the case of a high value of a PCF, a decimal can be omitted, in case of very low PCF more decimals than one decimal can be meaningful.

1.25 kg are rounded to 1.3 kg  $\rm CO_2$  equivalents; 1.24 kg are rounded to 1.2 kg  $\rm CO_2$  equivalents.

A PCF study shall **clearly specify** the declared unit of the system under study. The declared unit shall be **consistent with the goal and scope** of the PCF study [ISO 14067: 2018]. The primary purpose of a declared unit is to **provide a reference** to which the inputs and outputs are related. Therefore, the declared unit shall be clearly **defined and measurable**. An example of a **declared unit** is typically referring to the physical quantity of a product, for example "1 kg of liquid laundry detergent with 30 percent water content"

The declared unit for which the PCF of a product system is calculated is **1 kg of unpackaged** product at factory gate, regardless of its state (solid, liquid, gas), as its specific density is considered [BASF SE 2021]. If packaging is included (see 5.1.2), the declared unit is 1 kg of packaged product at factory gate.

TfS will consider specific guidance for the inclusion of packaging in the next revision of the guideline.

In all cases, a clear definition of the **declared unit** as basis for the PCF shall be disclosed. The calculations shall refer to the **declared unit** and shall be integrated in the deliverables when PCF data are exchanged between companies.

### 5.2 Calculation rules

### 5.2.1 Steps of PCF calculation

This chapter comprises the key calculation criteria to be followed while developing PCFs.

A PCF study in accordance with this document generally goes through the four phases of life cycle assessment, resulting in the following general steps:

- (i) Goal and scope definition: The declared unit shall be defined and all relevant activities and processes within the system boundaries identified. The system boundaries are outlined in chapter 5.1.2 and comprise all service, material and energy flows that become, make, and carry the product from raw material extraction to the factory gate.
- (ii) Creating the Life cycle inventory by collecting activity data: Activity data shall be collected for processes within the system boundaries (e.g. material input, energy inputs such as electricity, cooling and heating, purchased products and direct emissions). The applicable data requirements for the different types of activity data are described in chapter 5.2.8. See chapter 5.2.3 for details on which activities can be excluded from the collected data.
- (iii) Life cycle impact assessment:
- a. Calculating emissions: GHG emissions arising from a process shall then be calculated by multiplying the relevant activity data with its respective emission factor (CO<sub>2</sub>e per declared unit). The term activity data describes e.g. the input of materials, a process, a chemical reaction, a work up or purification step. Data types and emission factor sources are described in chapters 5.2.5 and 5.2.6.
- Additional steps can be required such as splitting emissions from multi-output processes or allocating them to different outputs. For guidance on such subjects see chapter 5.2.9.
- c. To allow for flexibility in applying accounting standards, calculations should be completed such that different allocation methods could be applied if needed. This ensures that different standard guidelines can be adhered to if required [Pathfinder Framework (PACT powered by WBCSD)], [BASF SE 2021].
- (iv) PCF consolidation: The PCF shall then be calculated summing up all GHG emissions.
- a. If the company produces the product in several different sites, bottom-up calculations for each production site using site-specific data, and if applicable, country-specific secondary data for processes not under the control of the reporting company, shall be performed. For communication purposes, the company may aggregate the sitespecific data into a weighted average based on the production volumes of the respective productions. If site-specific PCF data is averaged, this must be transparently stated. In addition, it will be reflected in a lower data quality score.

- b. In general, data collection should be as granular as possible, ideally from the specific processes involved in the production of the product under study. When process level data is not available, the data must be collected at plant or even site level, preferring plant level data to site level data. In these cases, emission factors from energy use or direct GHG emissions from a whole facility or site need to be attributed to the specific processes of the facility or site. This shall be done using a mass-based attribution approach. For this a break-down factor (BDF) is needed to attribute the GHG emissions from a facility or a site to the individual process. The BDF is calculated as a ratio of the production volume of the process [in tons] to the production volume of the facility or entire site [in tons.] Subsequently, the GHG emissions of the plant or site are multiplied by this BDF to result in process-level GHG emissions.
- (v) Documentation and reporting.

### 5.2.2 Temporal Scope

The time boundary of a PCF refers to the time period for which the PCF value is considered to be representative [ISO 14067: 2018]. The following time boundaries apply for the different types of data:

- Primary data used in the calculation of PCFs shall be as recent as practicable and not older than five years. The most recent full year (reporting- or calendar year) shall be applied as the time boundary for PCF calculations, if representative of an average year of production. For production years that are not continuous or irregular, production data may be averaged for a longer time period to reduce variability due to revisions, turnaround, or other atypical production conditions. When applying average production data in a PCF calculation, no more than the last three years of production (reporting- or calendar year) shall be averaged and used in a PCF calculation [BASF SE 2021], [Pathfinder Framework (PACT powered by WBCSD)].
- Secondary data used for all inputs and outputs should reflect the most recent activity data and/or the latest LCIs available. LCI data (e.g., from databases) used in the calculation of PCFs shall be as recent as practicable and not older than ten years [BASF SE 2021]. If older, appropriate, later proxies should be used instead. The data quality rating will be influenced by the choice of data.
- PCFs should be calculated on a regular basis to track improvements over time. However, this may pose a challenge for companies that rely on manual PCF calculations for products and who do not have an automated calculation approach. PCFs shall therefore have a maximum validity period of up to five years from the reference year of data collection if there have not been major changes to the production process (>20% impact from original PCF). Companies may update their PCF calculations on a more regular basis (e.g., annually). TfS decided that after five years or if the production process has changed significantly, PCF values are no longer considered representative and must be re-calculated. According to EN 15804 [EN 15804 - 2: 2019], an EPD is valid for 5 years as well, after which it must be re-verified and typically revised. If no changes are detected after 5 years, the PCF value can be renewed by a statement as well. Once a PCF has been revised, the revised version will take over the original PCF and be valid for 5 years. However, we recommend initiating a review process for a PCF after 3 years.



 The time boundary of the PCF calculation is the reference year. The PCF's reference year and date of calculation/publication shall always be disclosed alongside the PCF value.

### 5.2.3 Criteria to exclude certain activities (Cut-off)

In general, all processes, flows and activities, that are attributable to the product system shall be included in a PCF (see 5.1.2 on generally excluded and included activities) [BASF SE 2021] [ISO 14067: 2018]. The LCI data collection process shall aim for completeness. Where quantitative data are available, they shall be included. However, no undue effort should be spent on developing data of negligible significance concerning GHG emissions. If individual material or energy flows are found to be insignificant for the carbon footprint of a particular unit process, these may be excluded for practical reasons and shall be reported as data exclusions.

Cut-off criteria specify the amount of material or energy flow or the level of significance of GHG emissions associated with unit processes or the product system that may be excluded from a PCF study [BASF SE 2021]. Furthermore, cut-offs may become necessary in cases where **no data are available**, where elementary flows are very small (below quantification limit), or where the level of effort required to close **data gaps** and to achieve an acceptable result becomes prohibitive.

If no data are available, but elementary flows are significant, data gaps should be closed in accordance with chapters 5.2.6 and 5.2.8.

Several cut-off criteria are used in LCA practice to decide which inputs are to be included in the assessment, such as mass, energy, and environmental significance [BASF SE 2021]. Making the initial identification of inputs based on mass contribution alone may result in important inputs being omitted from the study. Accordingly, energy and environmental significance should also be used as cut-off criteria in this process.

### Requirements for PCF cut-off criteria

- All material inputs that have a cumulative total of at least 95% of the total mass inputs to the unit process shall be included. But we recommend covering 98% or more to remove potential uncertainties and increase the level of completeness [BASF SE 2021].
- 2. All energy inputs that have a cumulative total of at least 95% of total energy inputs to the unit process shall be included. To generate a PCF with higher quality by improving the completeness of the calculation, 98% of total energy inputs or more should be included.
- In cases where the input and influence on the PCF is unclear, an overall calculation should be made with generic figures to decide if a cut-off can be applied or not (iterative approach) [BASF SE 2021].

4. Input material flows that have a considerable upstream environmental footprint (e.g., precious metal like platinum group containing catalysts) should be considered in the PCF calculation, regardless of their relative contribution to the total mass of material flows, even if their mass input is < = 1% of the total mass. The PCF calculation should at minimum consider the loss of material (e.g., the loss of catalyst) and assign a PCF equal to the virgin material. If known, the efforts of recycling should be considered in addition. Otherwise known efforts, derived from other processes, can be used as a proxy.</p>

### 5.2.4 Standards used

This sectorial TfS quideline for chemicals follows the international standards ISO 14040:2006/AMD 1:2020 and ISO 14044:2006/AMD 2:2020 for Life cycle assessment. Derived from these generic standards, the guideline follows ISO 14067: 2018 for Product Carbon footprints (PCF). According to ISO 14067 [ISO 14067: 2018], the carbon footprint of a product is the "...sum of GHG emissions and GHG removals in a product system, expressed as CO<sub>a</sub> equivalents and based on a life cycle assessment using the single impact category of climate change." According to ISO 14067 [ISO 14067: 2018], a PCR is a "set of specific rules, requirements, and guidelines for carbon footprint of a product or partial carbon footprint of a product quantification and communication for one or more product categories." It also draws from other guidelines such as the **GHG Protocol** developed in recent years. The work of the Partnership for Carbon Transparency's Pathfinder Framework (hosted by WBCSD) and WBCSD Life Cycle Assessments guideline were considered as well. Generally, the guideline follows these standards and provides clarification and examples for the chemical industry.

To increase the consistency of PCF calculations along the value chain the following aligned prioritization hierarchy of guidelines shall be followed for PCF calculations:

- 1. PCR which was developed based on TfS Guideline.
- Product or sector specific guidelines based on ISO 14000 series (such as PCRs or Plastics Europe).
- 3. **TfS Guideline** if you do not have a PCR yet, the guideline can be used to calculate the PCF.
- 4. ISO 14067 standard [ISO 14067: 2018].
- Pathfinder Framework (PACT powered by WBCSD);
   GHG Protocol Product Standard [GHG Protocol Product Standard].
- Product Environmental Footprint Category rule (PEFCR) developed under the European Product Environmental Footprint initiative [EU PEF].

If different officially declared PCRs for the same product from different organizations exist, TfS will review them with an expert team and declare the "TfS accepted PCR". As a basis for the decision the correct application of the TfS guideline is first checked . TfS publishes and updates in every year a list of the "TfS accepted PCRs". In the case of sector-specific rules which are not officially declared as PCRs or PEFCRs, application shall also be justified and verified by TfS.

Table 5.2 TfS accepted PCR (list can be adopted after review of PCR by TfS experts)

Product system	Standard/Rationale followed
Steam crackers	[Plastics Europe - Steam Cracker Allocation [2017]]
C12-14 Fatty alcohols (oleo), methyl esters, refined oils, and crude oils from oil palm, refined- and crude oils from Coconut	[ERASM 2014]
Toluene diisocyanate (TDI), Methylene diphenyl diisocyanate (MDI)	[ISOPA 2012]
Chlorine (chlor-alkali process)	[EUROCHLOR 2022]

### 5.2.5 Data types and sources

Data can have different levels of quality. Every PCF calculation should be of the highest level of quality to be meaningful and applicable. High quality data are for example emissions data that are verified under a governmental scheme such as the EU-ETS. In a chemical reaction, several inputs are needed. Information about the inputs can be derived from different sources. The input from all sources shall be assessed with a quality rating system and data with the highest quality rates shall be used in the calculation of the PCF. For share of primary data and data quality rating, please refer to chapter 5.2.11.

Sources can be defined as:

### Primary data:

- Company-specific data refers to directly measured or collected data from one or more processes (processspecific data), from one or more facilities (facility- or plant-specific data) or from one or more sites (site -specific data) that are representative of the activities of the company (company is used as synonym of organization).
   To determine the level of representativeness a sampling procedure may be applied<sup>1</sup>.
- Primary data are defined as data from specific processes in the studied product's life cycle. They are collected for all processes under the ownership or control of the reporting company. Direct emissions data, emission factors and process activity data can be classified as primary data if they meet the definition.
- In general, primary, company -specific data should be collected and calculated as far as possible, i.e., at the highest level of granularity. This means that processspecific data is preferred over facility-specific data which is preferred over site-specific data.
- If only facility-specific or site-specific data of a company are available, they shall be collected or calculated and shall be representative of the facility or site for which they are collected.
- Facility or site-specific data shall then be broken down to the product level based on mass or other meaningful relations.
- Site-specific data should also be used for those unit processes that are commonly used for several processes, e.g. incineration or waste treatment. The overall consumption data should be calculated per service unit, e.g. kg CO<sub>2</sub>e per ton of waste incinerated. In addition, available information on specific emissions

in specific processes shall be considered (e.g.  $\rm SF_{_6}$  emissions from an incineration process of plasma that is used in the semiconductor industry).

Several standards prioritize the use of primary data, which is supported by this standard as well, if the data quality is high (see 5.2.11).

### Secondary data:

- Secondary data Defined as data that are not directly collected, measured, or calculated based on specific production data available for the company.
   Secondary data can include supplier and technological specific data derived from detailed data at plant/site level from market reports or patents, industry average data, or literature studies and can be an important and meaningful source for data included in PCF calculations.
- Secondary data includes industry averages, estimates based on literature studies, associations, published production data, government statistics, literature studies, engineering studies and patents and may also be based on financial data. It can contain proxy data generated by external expert judgement and other generic data. In addition, it can be sourced from a third party LCI database, open sources, PCF calculations, etc.
- It can be independently reviewed which increases
  the reliability and Data Quality Rating (DQR) score.
  Secondary data shall only be used for inputs and
  outputs where the collection of primary data is not
  practicable, or for processes of minor importance or
  where secondary for various reasons have a higher
  quality or fit better than primary data (e.g. association
  data for specific products).
- Secondary data can have the same level of quality as primary data, depending on the process of generation of the data, of meaningful fit to the data used, the level of aggregation etc.

### In case of data gaps

Data gaps exist when there is no primary or secondary data that is sufficiently representative of the given process in the product's life cycle. For most processes where data are missing, it should be possible to obtain sufficient information to provide a reasonable estimate. Therefore, there should be few, if any, data gaps. The data quality rating will indicate that there are data gaps existing which were filled by proxy data. The following sections give additional guidance on filling data gaps with proxy data and estimated data.

(1) Please see Appendix A of the GHG Protocol Corporate Standard for more information on sampling and sampling techniques.



### Proxy data

Proxy data are data from similar processes that are used as a stand-in for a specific process. Proxy data can be extrapolated, scaled up, or customized to represent the given process. Companies may customize proxy data to resemble the conditions of the studied process more closely in the product's life cycle if enough information exists to do so. Data can be customized to better match geographical. technological, or other metrics of the process. Identifying the critical inputs, outputs, and other metrics should be based on other relevant product inventories or other considerations (e.g., discussions with a stakeholder consultant) when product inventories do not exist.

Examples of proxy data include:

- Using data on polyethylene plastic processes when data on the specific plastic input (e.g., HDPE) is unknown. Depending on the specific assessment, the processes under study and the contribution to the overall PCF, using polyethylene data as a proxy for polypropylene might be sufficient as well.
- Adapting an electricity grid emission factor for one region to another region with a different generation mix.
- Customizing a process of another product to match the studied process, e.g. by changing the amount of material consumed to match a similar process in the studied product.

Table 5.3 Data hierarchy for energy and material inputs regarding primary, secondary and proxy data [Pathfinder Framework (PACT powered by WBCSD)]

Approach	Activity data source		Emission factor	source
	Energy¹	Material	Energy	Material
Best case	In-house/primary		For on-site production: In-house/primary For purchased electricity: Supplier-specific/ Renewable Electricity Certificates and Guarantees of Origin For other purchased energy: Supplier-specific	Supplier-specific (e.g. via Pathfinder Network)
Base case <sup>2</sup>	In-house/primary		Secondary databases	
Worst case <sup>3</sup>	In-house/secondary³ Proxy data		Proxy data and EEIO databases	

<sup>(1)</sup> Electricity, heating/cooling, steam

### **Estimated data**

(3) Financial data.

10

When a company cannot collect primary data or integrate meaningful secondary data or proxy data to fill a data gap, companies shall estimate the missing data to determine the significance of its contribution to the PCF result. If processes are determined to be insignificant based on estimated data, the process may be excluded from the inventory results (cut-off criteria). Criteria for determining insignificance are outlined in chapter 5.2.3 [GHG Protocol Product Standard]. If the data gap is significant and cannot be closed by the other types of data defined in this chapter, an estimation of the data shall be introduced. This should be done carefully under consideration of all knowledge of the data gap with a subsequent generation of estimated data. The estimated data shall be replaced by primary or secondary data as soon as possible in the update of the PCF. To assist with the data quality assessment, any assumptions made in filling data gaps, along with the anticipated effect on the product inventory results, should be documented [ISO 14067: 2018].

#### 5.2.6 Emission factor requirements and sources

Emission factors are the GHG emissions per unit of activity data, and they are multiplied by activity data to calculate GHG emissions. Emission factors may cover one type of

GHG (for example, CH,/liter of fuel) or they may include many gases in units of CO<sub>2</sub> equivalents. Emission factors can include a single process in a product's life cycle, or they can include multiple processes aggregated together. Life cycle emission factors that include emissions from all attributable upstream processes of a product are often called cradle-togate emission factors. Companies should understand which processes are included in the inventory's emission factors to ensure that all processes in the product's life cycle are accounted for in the data collection process.

Emission factors come from different sources and a distinction is made between primary and secondary emission factors:

Primary emission factors are emission factors calculated based on primary activity data for a process under a company's control or provided by a supplier for a process under their control.

**Secondary emissions factors** are derived from sources such as LCA databases, published product inventory reports, government agencies or industry associations. Secondary or default emission factors are based on secondary activity data. The source of secondary data must be specified in the report.

Emission factors shall always include all GHGs and be cradle-to-gate emission factors that include emissions from all attributable upstream processes of a product.

The following hierarchy shall be applied when selecting emission factors:

- 1. Where primary emission factors are available directly from raw material and energy suppliers, or internal processes. these shall be used. The quality of the supplier- or companyspecific emission factor is to be evaluated and checked for appropriateness (see below: data requirements on primary data or reference to appropriate chapter).
- 2. When using emission factors from utility companies, e.g., for electricity or steam (so-called market-based factors), it must be ensured that these are cradle-to-gate emission factors, including both, the emissions from combustion as well as the emissions from the provision of primary energy carriers. If the utility company cannot provide a life cycle emission factor, additional information such as the primary energy carriers used, and their respective shares needs to be disclosed. Based on this information, the upstream emissions from the provision of the energy carriers shall be calculated to complement the CO, emission factor from combustion to obtain a life cycle emission factor as described under 5.2.8 Activity data requirements. Additionally, the emission factors provided should include all GHGs but at least cover CO<sub>2</sub>, which is by far the largest contributor (>95%) to GHG emissions from combustion of primary fuels.
- 3. The utility providers should use either the efficiency or energy allocation approach when calculating emissions from Combined Heat and Power (CHP) installations plants, following the recommendations of the WBCSD accounting document which includes efficiency values by defaults to be used if needed [WBCSD Chemicals [2013]].
- 4. If primary emissions factors are not available, use secondary emission factors that are most suitable according to chapter 5.2.6. Among available data, use PCF values that are most representative and specific to the geography and technology used to produce the raw materials, utilities, and fuels. Only data from high quality and verified databases as listed below should be used as source of secondary data.

Additional requirements for the selection of secondary data for raw material apply as shown below. The following selection hierarchy shall be followed [BASF SE [2021]]:

- 1. If the production origin (region or country) and production technology of the supplied raw material is known, choose a regional or country/technology specific emission factor. A region can be the whole world, a group of several countries (e.g Europe) or a smaller area (e.g a group of states in the USA, a province in Canada) E.g Hydrogen liquid chlor-alkali electrolysis, membrane cell production in Europe.
- 2. If the production origin (region or country) of the supplied raw material is known, but the technology is not known, choose a regional or country-specific production mix, e.g Hydrogen liquid production in Europe.
- 3. If the production origin is not known, choose a regional or country-specific consumption mix based on the location of

- 4. If there is no regional or country-specific dataset available choose the same raw material from another country or region which is the most appropriate in terms of GHG emissions. E.g Hydrogen liquid chlor-alkali electrolysis, membrane cell in Europe for a supplier located in Brazil rather than using a global average value based on a high share of countries where the energy is mainly based on coal.
- 5. If the specific raw material is not available choose an appropriate proxy e.g., a chemical substance from the same chemical group.

Data quality of inbound and inter-site transports is based on primary data from a database for transport activities including emission factors of transport modes with a high quality.

In general, life cycle emission factors shall be sourced from and calculated based on data from verified sources such as listed below (non-exhaustive list):

- · Verified data from associations such as ISOPA, Plastics Europe, Fertilizer Europe, World Steel association etc.
- LCA databases such as GaBi (Sphera), Ecoinvent, Carbon Minds, Agribalyse, ELCD (PEF), IDEA database, etc.
- Official national emission factor databases such as US EPA, IEA, Defra, GREET etc.
- GLEC Framework [GLEC Framework] or DIN EN ISO 16258 for transportation.

If secondary emission factors are not available within the references listed above, other sources or proxy data may be used to fill in the missing emission factors. In any case, the source of secondary data or the employment of proxy data sources shall be reported. The extent to which secondary data is used shall be specified in relation to all GHG emissions by CO<sub>2</sub> equivalents.

The sources of secondary data shall be specified in the report. The attributes list requirements in Chapter 5.3. describe in detail, which attributes shall be reported for primary and secondary data as well as for the use of databases of secondary data.

### 5.2.7 Life Cycle Impact Assessment (LCIA)

A PCF represents the potential life cycle impact of a product on the environmental impact category of climate change. This impact category considers that different GHGs have different impact on climate change, expressed as their global warming potential (GWP) with the unit kg CO<sub>2</sub> equivalents (CO<sub>2</sub>e).

The basic equation to calculate GHG emissions (CO<sub>2</sub>e) for an activity data is:

Kg CO <sub>2</sub> e =	Activity data	Х	Emission factor	Х	GWP
	Amount of activity		(kg GHG/activity)		(kg CO <sub>2</sub> e/kg GHG)

For example, if the activity is the purchase of 5000 kg of methanol as a raw material and the supplier-specific emission factor is 0.80 kg CO<sub>o</sub>e/kg, then the GHG emissions for the activity  $5000 * 0.80 = 4000 \text{ kg CO}_{\circ}\text{e}$ .

11

your direct supplier, e.g Hydrogen liquid market in Europe.

<sup>(2)</sup> Prevalent approach in practice.



The basic equation to calculate CO<sub>2</sub>e for a direct emission is:

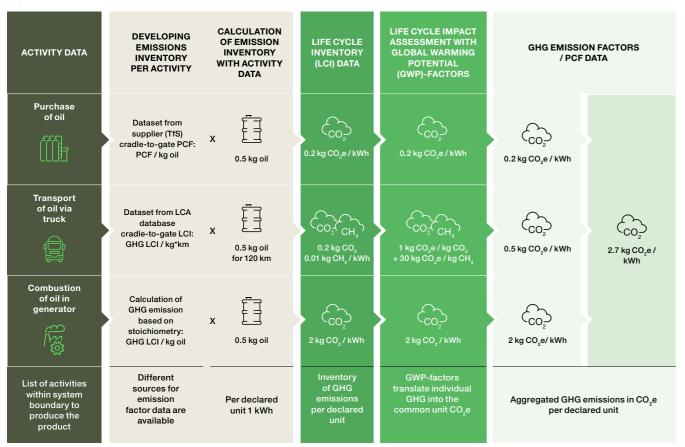
 $Kg CO_2e = Direct emission Data *GWP$ 

 $(unit) \qquad \qquad (unit) \, (kg \, GHG) \qquad \qquad (kg \, CO_2 e/kg \, GHG)$ 

### Formula 2

The types of emission factors needed depend on the types of activity data collected.

Figure 5.3 Types of data for PCF calculation on the example of production of electricity



In Figure 5.4, an example is described for the Chlor-alkali electrolysis gate-to-gate process data. The chlorine production weighted average of selected material and energy inputs and outputs are shown per kg chlorine. The values in the figure do not represent allocated but total in- and outputs of the average electrolysis process divided by the chlorine amount produced and just show

only some inputs. The allocation follows the generation of this GHG information. It is shown, how activity data and emission factors shall be introduced to generate a guideline compliant data set prior to allocation Euro Chlor [EUROCHLOR 2022]. Proxy secondary data for the PCF of input materials were extracted from Winnipeg [Winnipeg CO<sub>o</sub> Emission Factors].

Figure 5.4 Chlor-alkali electrolysis gate-to-gate process data of data for PCF calculation and transfer into a basic PCF prior to allocation

ACTIVITY DATA	DEVELOPING EMISSIONS INVENTORY PER ACTIVITY  CALCULATION OF EMISSION INVENTORY WITH ACTIVITY DATA	LIFE CYCLE INVENTORY	LIFE CYCLE IMPACT ASSESSMENT WITH GLOBAL WARMING POTENTIAL (GWP)-FACTORS	GHG EMISSION / PCF D/	
Using grid electricity	Calculation of GHG emission based on data X 2.36 kWh from grid: GHG LCI / kWh	0.395 kg CO <sub>2</sub> /kWh	CO <sub>2</sub> 1 kg CO <sub>2</sub> e / kg CO <sub>2</sub>	0.93 kg CO <sub>2</sub> e / kg	
Purchase of salt	Dataset from supplier (TfS) X 2.15 kg salt cradle-to-gate PCF: PCF / kg oil	0.2 kg CO <sub>2</sub> /kg	CO <sub>2</sub> 1 kg CO <sub>2</sub> e / kg CO <sub>2</sub>	0.43 kg CO <sub>2</sub> e / kg	1.4 kg CO <sub>2</sub> e / kg
Purchase of sulphuric acid	Dataset from LCA database X 0.01 kg cradle-to-gate LCI: GHG LCI / kg	0.14 kg CO <sub>2</sub> /kg	CO <sub>2</sub> 1 kg CO <sub>2</sub> e / kg CO <sub>2</sub>	0.001 kg CO <sub>2</sub> e / kg	
List of activities within system boundary to produce the product	Different sources for emission factor data are available  Per declared unit 1 kg Chlorine	Inventory of GHG emissions per declared unit	GWP-factors translate individual GHG into the common unit CO <sub>2</sub> e	Aggregated GHG em	

The PCF calculation consists of the sum of each GHG released and removed from the product system and application of allocation rules when necessary (see chapters 5.2.9 and 5.2.10).

The GHGs that shall be accounted for are identified within the GHG Protocol titled "Required Greenhouse Gases in Inventories: Accounting and Reporting Standard Amendment". The list includes Carbon dioxide ( $CO_2$ ), Methane ( $CH_3$ ), Nitrous oxide ( $N_2O$ ), Hydrofluorocarbons (HFCs), Perfluorinated compounds, Sulphur hexafluoride ( $SF_9$ ), Nitrogen triflouride ( $NF_3$ ), Perfluorocarbons (PFCs), Fluorinated ethers (HFEs), Perfluoropolyethers (e.g. PFPEs), Chlorofluorocarbon (CFCs) and Hydrochlorofluorocarbon (HCFCs). The GHG emissions shall be aggregated as  $CO_2$ -equivalents and should not be reported separately for individual gases.

The 100-year GWP characterization factors (GWP100y) according to the Intergovernmental Panel on Climate Change (IPCC) shall be used in the PCF calculations, based on the IPCC's Sixth Assessment Report (AR6). These factors include Climate carbon response for non-CO $_2$  gases. If in future there will be updates, TfS will update the guideline accordingly to follow the latest version.

The AR 6 GWP-100 characterization factors **shall be extracted in priority from Table 7.15** of Chapter 7 of the IPCC AR6 Climate Change 2021 Physical Science Basis. This table includes the chemical effects of  $\mathrm{CH_4}$  and  $\mathrm{N_2O}$  [IPCC 2021-The Physical Science].

The AR 6 GWP-100 characterization factors for the substances that are not listed in the Table 7.15 shall be extracted from **Table 7.SM.7** in the Chapter 7 Supplementary Materials of the AR6 Climate Change 2021 Physical Science Basis [IPCC 2021-The Supplementary Material].

The 100-year GWP characterization factors according to the IPCC's Fifth Assessment Report (AR5), Appendix 8.A (Lifetimes, Radiative Efficiencies and Metric Values) may be used in 2022 during the transition period [IPCC 2013- The Physical Science].

The PCF report shall disclose which IPCC Assessment Report basis is used.

### 5.2.8 Activity data requirements

Activity data describe specific applications and uses of materials, energies, services etc. In an LCA the description of activities within a system boundary is needed to generate mass flows of materials uses, energy uses, etc. The amounts of the activities are later linked with life cycle inventories to calculate the contribution of this activity to the PCF of the whole product.



### 5.2.8.1 Electricity and thermal energy

This chapter provides guidance on how to account for the emissions associated with the use of electricity and thermal energy such as steam, heat and cooling.

The GHG emissions associated with the use of energy should include

- Upstream emissions from the energy supply system (e.g. the mining and transport of fuel to the energy generator or the growing and processing of biomass for use as a fuel).
- GHG emissions during generation of electricity or thermal energy, including losses during transmission and distribution.
- Downstream emissions (e.g. the treatment of waste as ashes arising from the operation of coal fired power plants)

For sources of emission factors see chapter 5.2.6. If sources such as IEA or EPA are used, it shall be ensured that emissions associated with upstream activities are also included.

A company may purchase primary energy carriers such as natural gas, oil or coal either as a raw material for further material processing or as fuel to generate energy. The upstream emissions from activity to provide these primary energy carriers shall be estimated as described in chapter 5.2.8.2. Bay materials.

### Thermal energy: Steam, heat and cooling systems

Companies shall report emissions from the purchase and use of these energy products the same as for electricity: according to a location-based and market-based method if the contractual instruments used meet the Scope 2 Quality Criteria as appropriate for gas transactions. These may be the same total where direct line transfers of energy are used [GHG Protocol Scope 2 Standard].

### Self-generated thermal energy

If the energy is internally generated (e.g. on site) and consumed for the production of the studied product, the primary data of the energy generation system shall be used to calculate the PCF of the product. Primary data for both, activity data and direct emissions shall be collected via a bottom-up approach.

Thermal energy may also be generated as a co-product of a chemical processes (e.g. excess steam). See chapter 5.2.9 for further guidance on how to account for emissions from energy and other co-products.

### Purchased thermal energy

If the reporting company purchases thermal energy, GHG emission factors from a supplier-specific energy product shall be used (market-based approach).

A market-based method reflects emissions from electricity that companies have purposefully chosen (or their lack of choice). It derives emission factors from contractual instruments, which include any type of contract between two parties for the sale and purchase of energy bundled with attributes about the energy generation, or for unbundled attribute claims.

If the utility provider is not able to provide a life cycle based GHG emission factor for the energy product but only the CO<sub>2</sub>e emission factor from direct emissions (e.g. combustion), the upstream emissions for the fuels that go into the energy production need to be added. In this case, the energy provider needs to provide information on the primary energy carriers used and their share. The GHG emission factors shall be rated with a DQR assessment following this standard.

### **Electricity**

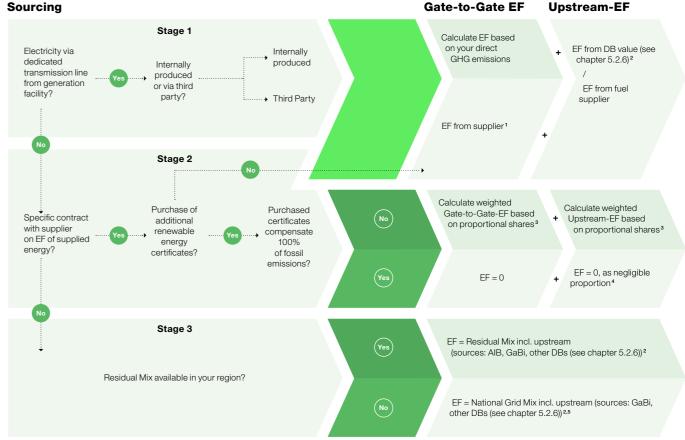
For the use in the PCF calculation organizations should generally calculate the emissions of electricity following the market-based approach (as described in the GHG Protocol Scope 2 Guidance). The electricity accounting approach used should be addressed in the PCF reporting. Please follow the decision tree in Figure 5.5 to determine your options on GHG emissions of procured electricity. As stated above the total GHG emission factor should include GHG emissions during generation of the electricity (gate-to-gate) and upstream emissions from the primary energy supply system. For convenience it is possible to add both factors to result a total GHG factor if both refer to the same energy unit. The decision tree is divided into the three stages (which are additionally explained below in more detail):

- Stage 1: Electricity via a dedicated transmission line (market-based).
- Stage 2: Electricity from the grid (location-based) or specific contract with supplier on energy mix (market-based).
- Stage 3: Residual Mix (no specific contract with supplier on energy mix or no specific data available).

Start in the top left corner of stage 1. Exception: If your company has sold energy attribute certificates for received electricity via a contractual instrument to a third party, start in stage 3 (see Figure 5.5).

Gate-to-gate emission factors consider emissions within the company boundary excluding all upstream emissions.

Figure 5.5 Decision-Tree on selection of proper emission factors for externally sourced electricity



- (1) If the Emission Factor (EF) from supplier is not available, directly move to stage 3.
- (2) If no access to Upstream-EF data, please apply 20% of the IEA value instead and add it to the Gate-to-Gate-EF.
- (3) After receiving the individual energy mix from your supplier, multiply the EFs corresponding to their energy source with their proportional share of the energy mix while also taking the partly compensated fossil emissions by purchased certificates into account (e.g.: energy mix: 20% renewable energy (RE), 80% fossil energy (FE); purchased certificates: an amount to compensate 50% of sosil emissions
- $= EF_{Weighted} = 0.2 \times EF_{RE} + 0.8 \times 0.5 \times EF_{FE} + 0.8 \times 0.5 \times 0).$
- (4) If impact lies within the cut-off range (s. chapter 5.2.3), apply EF = 0. Otherwise, please use DB value (GaBi or other DBs (see chapter 5.2.6)).
- (5) Alternatively, IEA-Data can be implemented if additional Upstream-EFs from DBs (GaBi or other DBs (see chapter 5.2.6)) are added.

## Stage 1: Check if electricity is via a dedicated transmission line from the generation facility

### Determining the gate-to-gate emission factor

If there is a dedicated transmission line between the organization and the electricity generation plant and no certificates (also known as contractual instruments) for that consumed electricity have been sold to a third party, GHG emission factors from the supplier-specific electricity shall be used.

- If the electricity is internally generated (e.g. on-site generated electricity) primary data of the electricity generation system shall be used to calculate the PCF of the product.
- If the electricity is provided by a third party, a GHG emission factor obtained from the third party may be used.

If there is a dedicated transmission line between the organization and the electricity generation plant and energy attribute certificates have been sold by contractual instruments to a third party, then the organization must start in stage 3 of the decision tree.

### $\label{eq:decomposition} \textbf{Determining the upstream emission factor}$

Additional upstream GHG emissions (e. g. from mining and transport of fuels to the electricity generation facility) can either be requested from the suppliers of fuel or electricity or calculated from database values (suitable databases see chapter 5.2.6). If the organization has internally produced electricity and decides to calculate upstream GHG emissions from database values, the fuel consumption per unit of electricity produced serves as a basis. In case of electricity from third parties the composition of the electricity mix is required for calculation.

### Stage 2: Electricity from the grid (specific contract with supplier on energy mix)

### Determining the gate-to-gate emission factor

If the organization has a specific contract with an electricity supplier regarding electricity with a certain GHG emission factor and no further renewable energy attribute certificates are purchased, then the organization shall use GHG emissions from a supplier-specific electricity product.



In the case that further renewable energy certificates are purchased, the organization must check if they are sufficient to cover the fossil emissions of the obtained electricity. If not, then a proportional gate-to-gate emission factor for the electricity must be calculated based on the remaining share that is not covered by the certificates. If the certificates compensate the fossil emissions, the gate-to-gate emission factor can be set to zero.

Please note that via contract the electricity supplier must guarantee that their product is tracked to ensure that no double counting of renewable electricity occurs.

#### Determining the upstream emission factor

Additional upstream GHG emissions (e. g. from mining and transport of fuels to the electricity generation facility) can either be requested from the suppliers of electricity or calculated from database values (suitable databases see chapter 5.2.6). If the organization decides to calculate upstream GHG emissions from database values, the composition of the electricity mix is required for calculation.

In the case that further renewable energy certificates are purchased, the organization must check if they are sufficient to cover the fossil emissions of the obtained electricity. If not, a proportional upstream emission factor for the electricity must be calculated based on the remaining share that is not covered by the certificates. If the certificates compensate the fossil emissions in the gate-to-gate factor, the organization should determine the upstream emissions of the applied renewable energy type by calculation from database values. The upstream emissions may be neglected if they are insignificant and thus fall under the cut-off criteria (see chapter 5.2.3). To verify that, primary data should be used. If they are not available, secondary data information may be helpful for verification of the cut-off.

# Stage 3: Residual Mix (no specific contract with supplier on energy mix or specific data is not available)

When information on supplier specific electricity is not available or renewable attribute energy certificates have been sold to a third party, a residual GHG emission factor should be used (market-based approach). This factor represents the emissions that remain after certificates, contracts, and supplier-specific factors have been claimed and removed from the calculation. Organizations should check databases (see chapter 5.2.6) for residual mixes available for their region of operation. Database values are preferred if they cover a cradle-to-gate scope. Alternatively, organizations operating in Europe can use residual mixes from sources such as AIB [AIB 2021- European Residual Mix to determine their gate-to-gate emission factors. If this source is used, the upstream emission factors must be calculated based on the composition of the electricity mix using database values for the fuels. If AIB RES mix are used, upstream emissions for electricity should be calculated based on the fuels used. Companies operating in other regions should check if residual mix data is available (e. g. for certain US regions residual mixes are published, cf. [Green-e 2021-Residual Mix Emission Rate].

If no residual mix data is available, then as a last quality option according to the GHG Protocol Scope 2 Guidance [GHG Protocol Scope 2 Standard], national grid mixes can be applied. Organizations should check databases

(see chapter 5.2.6) for emission factors covering a cradleto-gate boundary. If no database values are available, organizations can use IEA data as gate-to-gate emission factors. If that route is chosen it is mandatory to calculate upstream emission factors based on the composition of the grid mix applying database values for the fuels.

#### Further notes on renewable electricity

The Renewable Energy Directive [EC-Renewable Energy Directive] defines renewable energy or "green" energy RES-E as: "...energy from renewable non-fossil sources, namely wind, solar, aerothermal, geothermal, thermal, hydrothermal and ocean energy, hydropower, biomass, landfill gas, sewage treatment plant gas and biogases".

Importantly, double-counting must be avoided. According to ISO 14067 [ISO 14067: 2018], no double-counting occurs:

- Where the process that used the electricity and no other process may claim the generator specific GHG emission factors for that electricity.
- Where the generator-specific electricity production does not influence the GHG emission factors of any other process or organization [ISO 14067: 2018].

The purchase and use of green electricity can be considered in the market-based emission factor provided that the criteria in ISO 14067 Chapter 6.4.9.4.4 are met ISO 14067: 2018).

If a unit is running with 20% certificates of 100% renewable energy, the total production can be claimed as being renewable by 20%. Alternatively, a mass balance approach can be applied to renewable or decarbonized electricity. In this case, the same principles as the mass balance chain of custody (chapter 5.2.10.5) for biomass can be applied. Renewable energy purchased for specific products may be applied to those specific products.

Offsets shall not be used in the calculation of renewable energy.

### Additional notes:

- If processes within the system under study are in small island developing states (SIDS, as defined by the United Nations), the PCF or the Cradle-to-Gate PCF may additionally be quantified using contractual instruments for such processes, irrespective of grid inter-connectivity.
- Contractual instruments are any type of contract between two parties for the sale and purchase of energy bundled with attributes about the energy generation, or for unbundled attribute claims.
   EXAMPLE: Contractual instruments can include energy attribute certificates, renewable energy certificates (RECs), guarantee of origin (GoOs) or green energy certificates.
- Characteristics of a generator should include the registered name of the facility, the name of the owners, the nature of the energy generated, the generation capacity and the renewable energy supplied. Additional characteristics can be added to describe the electricity generation.

### 5.2.8.2 Raw materials

Raw materials are defined as materials that are purchased and used to produce a product. They can be of primary or secondary origin. Secondary materials include for example recycled material. ISO 14040 [ISO 14040: 2006], see chapter 5.2.8.4). Primary raw materials are often named "virgin" materials.

According to the Pathfinder Framework [Pathfinder Framework (PACT powered by WBCSD)], raw materials can be:

- Extracted directly by the company, e.g. mining activities or agricultural production.
- Sourced by external suppliers.
- Toll manufactured.
- · Coming from recycling processes.

Chemical products are often based on raw materials that are derived from oil and its derivates. Raw materials supplied to a machine or processing plant are defined as feedstocks.

The PCF calculation shall consider the full upstream life cycles of raw materials; from raw material acquisition and pre-processing or direct generation from natural resources (e.g. mining) to the factory gate. It shall also include disposal of wastes generated during raw material production.

According to Pathfinder Framework [Pathfinder Framework (PACT powered by WBCSD)], material acquisition refers to the extraction of resources from the environment needed to create a product. Pre-processing refers to the refining of all the acquired natural and biogenic resources so they can be used in a production facility. Transportation to and from the sites of resource extraction, pre-processing facilities and production facilities shall also be included.

### Information on purchased raw materials and raw materials used in a chemical reaction

In chemical reactions, raw materials can be purchased or used from different sites or different plants within a site.

Production network ratios of chemical products and consumption mixes of raw materials should be defined as a basis for PCF calculations. The relationships between products from different sources should be documented with a bill of materials (BOM) from a reporting system. Intracompany relations between all involved sites of a company can be integrated in a network of information. Representative averages of the production network ratios (percentage rate) should be generated by solving and eliminating inter-company relations. Consolidated BOM will be used for the calculations. Ratios are available for all raw materials needed in one company based on a Supply-Demand-Balance for each production/site/plant and company information. To build averages of inputs of the same raw material from different sources, a mass weighting approach linked with the PCF of the different raw materials sourced shall be used.

### The average calculation can be based on:

- External source (purchased from external supplier):
  - Raw material is procured from an external supplier.
- All purchased raw material comes with a PCF. PCF information needs to be obtained either by supplier specific PCF provided with the raw materials or by secondary data for the raw material (see 5.2.5 on requirements for primary and secondary data and 5.2.6 on requirements for emission factors).
- For various suppliers of a raw material, PCF of raw materials should be averaged by amount of purchased volumes. As an alternative, supplier-specific raw materials may be segregated to specific product lines with documented justification.
- Company source:
- Product is produced per another BOM at the same company.
- Inter-company transferred product: product is sourced per a BOM from another internal site or even plant.
- Mixed source:
- Product is produced in another BOM at the same internal site/plant, and/or product is sourced from another site/plant of the company, and product is procured from an external provider [BASF SE [2021]].

The equation in section 5.2.7 shows a basic equation to calculate GHG emissions ( $CO_q$ e) from activity data.

Data used for raw materials can be primary or secondary data (see chapter 5.2.5). Further requirements on emission factors can be found in Chapter 5.2.6.

There are no minimum data quality requirements (see chapter 5.2.11) for raw materials currently to accommodate the need for a transition time for capability development in the supply chains. It is desirable for TfS or member companies to implement minimum data quality requirements in the future.

### 5.2.8.3 Transport

GHG emissions from transportation often have a minor impact on the PCF of a chemical product. However, they shall be considered and checked if important to the PCF by an iterative process (see also cut-off criteria, chapter 5.2.3).

The following transportation activities shall be included in a cradle-to-gate PCF:

- Transportation in the supply chain, for example the transportation of raw materials to the company site, or transportation of a raw material from a tier 2 supplier to a tier 1 supplier (if not already considered).
- Only if the contribution to the overall PCF is significant (see chapter 5.2.3), in-bound transportation as e.g., transportation to an internal storage location as part of a company's direct activities should be considered.
- The transportation of an intermediate product from one production site to another shall be considered if relevant according to the cut-off criteria.



GHG emissions of outbound transportation shall not be included in the cradle-to-gate PCF but calculated and reported separately if requested by customers.

In general, the GHG emissions relating to the entire fuel life cycle (i.e., well-to-wheel)¹ shall be considered in the calculation of emissions from transportation.

Transports can either be carried out directly by the reporting company e.g., in company-owned or leased vehicles, or by external transport service providers. As such, the method used to calculate product-related transport emissions is very much dependent on the availability of information such as fuel consumption, distance covered, mode of transport or load specifics.

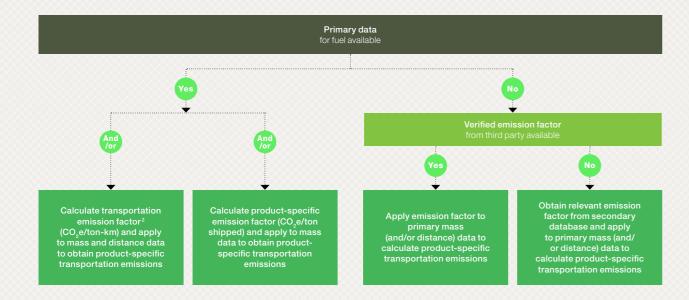
The following paragraphs provide guidance on how to calculate transportation emissions depending on the type of data available (see also Figure 5.6), [Pathfinder Framework (PACT powered by WBCSD)]. This guidance is not available anymore in the updated version of the Pathfinder Framework.

 If available, primary data on fuel usage should be used to calculate product-related transport emissions, based on actual transportation mode, distance and vehicle load.
 The fuel consumption data should cover the full round trip that is, include all fuel associated with full, partially loaded, and empty trips, when relevant. Allocation of

- these emissions shall be based on the mass of the product. In cases where transport is volume limited (full freight's mass is lower than the truck's load capacity) allocation shall be based on volume.
- Where primary data are not available, but data on productspecific transportation emissions has been shared by the third party operating the transportation, this data should be used and included in the PCF calculation.
- 3. When a company has neither primary data on fuel usage nor access to product-specific transportation emissions, primary data on mass and most suitable distance shall be used for the calculation of emissions. The relevant emission factor per type of transportation (expressed in CO<sub>2</sub>e per ton-km) e.g., provided by the transport service provider, should be applied to this data to calculate product specific emissions. If no emission factor is available, relevant secondary databases shall be consulted to obtain the necessary emission factor (see section 5.2.6 for suitable databases or [GLEC Framework]).

NOTE: Aircraft GHG emissions have additional climate impacts under certain circumstances at high altitudes because of physical and chemical reactions with the atmosphere. For more information on GHG emissions from aircraft, see the IPCC Guidelines for National Greenhouse Gas Inventories and the IPCC Special Report on Aviation.

Figure 5.6 Steps for calculating product transportation emissions [Pathfinder Framework (PACT powered by WBCSD)]



(1) Well-to-wheel includes the GHG emissions related to fuel production, distribution, and combustions. (2) Emission factors are always per transportation mode and type.

18

### Assessment of impacts from transport: example truck transport

Datasets for truck transport are per tkm (ton\*km) expressing the environmental impact for 1 ton (t) of product that is transported for 1km in a truck with a certain load. The transport payload (= maximum mass allowed) is indicated in the dataset. For example, a truck of 28-32 t has a payload of 22 t: the LCA dataset for 1 tkm (fully loaded) expresses the environmental impact for 1 t of product that is transported for 1km within a 22 t loaded truck. The transport emissions are allocated based on the transported product's mass and you get only a share of 1/22 of the truck's full emissions. When the load transported is lower than the maximum load capacity (e.g. 10 t), the environmental impact for 1 t of product is affected in two ways. First, the truck has less fuel consumption per total load transported (which is not considered for simplification reasons) and second, its environmental impact is allocated by the load transported (e.g., 1/10 t). When a full freight's mass is lower than the truck's load capacity (e.g. 10 t), the transport of the product may be considered volume limited. In this case, the environmental impact shall be calculated using the real mass loaded. If it is known that empty return transports are the case, the impact of the transportation emission from the round trip shall be considered and attributed to the transported product. For the empty return transport. a reduced emission factor can be considered compared to the full payload.

Based on the assumption of an average load factor of 0.5 net-tons per gross ton can be considered. It can be concluded that the share of empty vehicle-km in long distance transport is still significantly higher for rail compared to road transport. The additional empty vehicle-km for railways can be partly attributed to characteristics of the transported goods.

Therefore, we presume smaller differences for bulk and volume goods and make the following assumptions:

- The full load is achieved for the loaded vehicle-km with bulk goods. Additional empty vehicle-km is estimated in the range of 60% the maximum load for road and 80% of the maximum load for rail transport.
- The weight related load factor for the loaded vehicle-km with volume goods is estimated in the range of 30% of the maximum load for road and rail transport. The empty trip factor is estimated to be 10% for road transport and 20% for rail transport related to the maximum load. These assumptions consider the higher flexibility of road transport as well as the general suitability of the carrier for other goods on the return transport.

EcoTransIT World offers an emission calculator for GHG and exhaust emissions in compliance with EN 16258 and the GLEC Framework [EcoTransIT- Emission Calculator for GHG Emissions].

ISO 14083 that is under development will give further guidance for transportation. All assumptions and cut-offs considering transportation shall be reported. Furthermore, the Global Logistics Emissions Council (GLEC) developed the GLEC Framework, a globally recognized methodology for harmonized calculation and reporting of the logistics GHG footprint across the multi-modal supply chain may be applied [Global Logistics Emissions Council (GLEC)].

### 5.2.8.4 Waste treatment and recycling

Manufacturing of chemical products often involves the generation of waste materials, including solids, liquids, gases, and wastewater.

A waste is any substance or object which the holder discards or intends to discard per European Waste Framework Directive [EU Waste Framework Directive]. Waste has no economic value.

A co-product is a product that is produced in a multi-output process incidentally to the production of products that are intendedly produced and have the highest economic value in such a process. Co-products have an economic value and shall be considered for PCF calculations. See chapter 5.2.9 for guidance on how to account for valuable co-products.

This chapter provides guidance on calculating the burdens and benefits of waste treatment and recycling processes. This is relevant to the PCF calculation in three cases:

- Treatment of wastes generated from operations related to product manufacturing.
- The usage of energy which is recovered from waste incineration for product manufacturing.
- The usage of recycled secondary materials in the manufacturing of the product.
- Preparatory steps and supporting activities for all waste treatment- like collection, transportation, sorting, dismantling, or shredding- shall be considered and included in the PCF calculation following the guideline as described below.

Due to the cradle-to-gate boundary of the PCF calculation within this guideline, emissions from the use and end-of-life stage of the product itself shall not be included in the PCF calculation. If materials are used for the product as raw materials in a circular approach, they shall be considered following the relevant chapters in this guideline.

For the consideration of biogenic carbon please refer to chapter 5.2.10.1

### **Emission factor sources:**

- Whenever possible, companies should use waste treatment emission factors based on primary data.
- If the waste is treated by the company who generates it, the emission factor shall be calculated based on internal primary data.
- If the waste is sent to a third party for treatment, the treatment provider shall calculate their waste treatment emissions, develop emission factors, and verify and communicate these to the company who has generated the waste. The emission factors from the third-party treatment shall be calculated based on the TfS approach.

(1) Refer to Waste Framework Directive (2008/98/EC) for further definition requirements of by-products



- If primary emission factors cannot be obtained, secondary emission factors shall be used in the following hierarchy:
- Emission factors shall be estimated based on available information on the waste composition and process technology and parameters of the applied treatment technology. The calculation shall be based on the TfS approach.
- If this is not possible, emission factors should be derived from accepted secondary databases (chapter 5.2.6).
- In the case of no data is available, some proposals to develop proxies for landfilling and Wastewater treatment are shown in the appendix.

### Guidance on calculating emission factors for waste treatment and disposal

Emissions from the treatment of non-recycled waste generated during production shall be allocated to the main product or co-products and therefore shall be reflected in the PCF. Since waste is considered an output without economic value, no production emissions are allocated to the actual waste generated during production.

Typical waste treatment operations include disposal activities such as:

- Landfill.
- · Wastewater treatment.
- Incineration without energy recovery (see example 1).
- · Hazardous waste treatment.

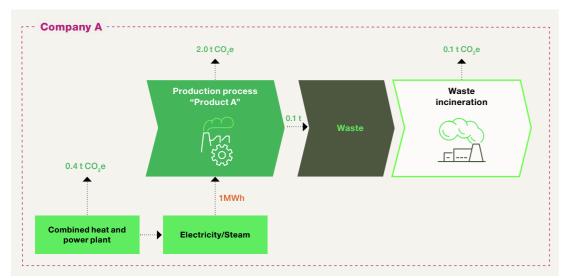
In some cases, different types of waste streams are co-treated in a single waste treatment facility, for example in the case of co-incineration of high and low calorific value waste streams or wastewater treatment for wastewater streams with different compositions. Such a waste treatment processes are multi-functional, regardless of whether it includes energy recovery. If data is available, then the impact of the incineration process shall be allocated to the different waste types following the allocation hierarchy for multi-functional processes as described in chapter 5.2.9.

### Example 1: Waste incineration without energy recovery

Waste from the manufacturing process of product A is incinerated without energy recovery (either on site or by a third party).

The impact of the incineration process should be calculated or estimated based on the requirements outlined in this guideline. The resulting emission factor shall be allocated to the PCF of product A.

Figure 5.7 Waste incineration without energy recovery and without use of the energy



 $PCF_{Product A} = 2.0 \text{ t CO}_{2}e/\text{ t} + 0.4 \text{ t CO}_{2}e/\text{ t} + 0.1 \text{ t CO}_{2}e/\text{ t} = 2.5 \text{ t CO}_{2}e/\text{ t}$ 

### Guidance for calculating emission factors for waste treatment with energy recovery

"Energy recovery from waste is the conversion of non-recyclable waste materials into usable energy such as heat or electricity, through a variety of processes, including combustion and other processes to recover energy. This process is often called "waste to energy" [EPA].

The impact of waste treatment with energy recovery shall be included in the product life cycle inventory and system boundary following the calculation approach outlined in this sub-chapter.

Material recycling processes are such processes that derive a secondary material from a waste material which is further used as material for manufacturing of products. Such processes are for example chemical recycling through pyrolization, distillation or mechanical recycling. Guidance on the calculation approach for material recycling can be found below.

Material recycling and waste treatment with energy recovery are considered separate and not equal. To reduce the emission of GHGs, the chemical industry should strive to keep carbon in a material loop. This is primarily achieved through the reduction of waste generation and material recycling of remaining waste. The impact attribution approach should be designed to incentive both.

Incineration is the least favorable solution because it is a final disposal. The different available calculation approaches regarding waste treatment with energy recovery have been discussed among TfS group members and no consensus has been reached so far. This document in the current state discusses three approaches, which are described

with their pros and cons below (Table 5.4). One of the three allocation approaches shall be followed. The choice shall be documented and communicated through the additional information of the PCF.

The discussion to select the most appropriate guidance in this chapter will be continued inviting additional stakeholders to contribute. The guideline will be updated accordingly to reflect changes and consensus. TfS also encourages the development of targeted solutions for such cases through among others, product category rules.

### Energy recovery within the system boundaries of a product

If all processes related to energy recovery from waste are included in the system boundary, an allocation is not required, or all allocation approaches lead to the same result. This is the case if the energy generated is directly used in the process of the studied product. The impact of the waste incineration shall be included in the PCF (see Example 2). This closed loop recycling means that the direct recycled energy has no additional environmental impact (= 0). The same applies for material recycling within the system boundaries, as described in the subchapter below.

### Example 2: Waste incineration with energy recovery within the system boundaries

Waste from the manufacturing process of product A is incinerated with energy recovery on-site and under operational control. The recovered energy is used in the production process of Product A. Since the recovered energy is used within the system boundaries of Product A, no allocation is needed. All  $\mathrm{CO}_2$ e emissions from the process shall be attributed to Product A.

Figure 5.8 Waste incineration with energy recovery within the system boundaries of the company



PCF<sub>Product A</sub> = 2.0 t CO<sub>2</sub>e/t + 0.2 t CO<sub>2</sub>e/t + 0.4 t CO<sub>2</sub>e/t = 2.6 t CO<sub>2</sub>e/t



### Energy recovery outside of the system boundaries of a product

Waste material is part of the life cycle of a product system. It can be treated with energy recovery and this energy can be used in additional product systems. This creates the need to split the impact of the treatment process and identify the part of the impact to be added to each product system.

### The following general rules shall apply:

- Whenever applicable and possible, process subdivision shall be used to divide common processes to avoid the need for allocation [GHG Protocol Product Standard (2011)].
- For waste treatment with energy recovery, whenever available, allocation methods in line with published and accepted product category rules (PCR) shall be applied.
- If none of the above apply, either of the three allocation approaches described below shall be applied. The choice shall be documented and communicated through the additional information of the PCF.

The following table describes the three different approaches and discusses its pros and cons. Any of the three methods can be used until further updates following ongoing discussions through TfS.

Table 5.4 Overview of different assessment approaches

	Cut-off approach also known as recycled content approach	Reverse Cut-off approach also known as waste allocation	Substitution
Description	"Energy producer takes control" All burden allocated to generated energy	"Polluter pays"  All burden allocated to waste generation process	"Market implications considered" Emissions from incineration reduced by credit for substituted energy
Who carries the burden?	Energy user(s)	Waste generator	Energy user(s) and waste generator
Who receives the benefit?	Waste generator	Energy user	Energy user(s) and waste generator
Pros	<ul> <li>+ Incentivizes waste treatment with energy recovery compared to without</li> <li>+ In alignment with GHG Protocol and WBCSD Pathfinder</li> <li>+ Simple to apply</li> </ul>	<ul> <li>+ Incentivizes waste reduction</li> <li>+ Incentivizes energy recovery from waste treatment</li> <li>+ Simple to apply</li> <li>+ Simple data exchange (waste generator provides waste data for calculation and receives emission factor)</li> </ul>	+ Incentivizes waste treatment with energy recovery compared to without  + GHG & ISO conform  + Commonly implemented in LCA databases  + Incentivizes waste reduction if more renewable energy is available
Cons	<ul> <li>No incentive for material recycling compared to energy recovery</li> <li>No incentive to reduce waste</li> <li>No incentive to use energy compared to renewable energy (Higher emission factors compared to best technology)</li> <li>Some LCA database need to be adjusted</li> </ul>	<ul> <li>Deviates from GHG Protocol</li> <li>No difference in energy emission factor compared to renewable sources</li> <li>Lower incentive for energy reduction</li> <li>Some LCA database need to be adjusted</li> </ul>	<ul> <li>Result depends strongly on selected comparative system for substitution</li> <li>Complex data exchange data for comparative solution required (market data) and agreed by energy user and waste provider</li> </ul>
Link to/ Implications for corporate GHG emissions reporting	In line with corporate GHG reporting	Corporate reporting has to be adjusted	Substituted emissions need to be reported separately

### Following the cut-off approach (also known as recycled content approach):

- The impact of preparatory steps and supporting activities such as collection, transportation, sorting, dismantling, or shredding shall be added to the inventory results of the product system generating the waste.
- The waste input to the energy recovery process shall be treated as free of burdens. Burdens or credits associated with material from previous or subsequent life cycles are not considered i.e., are "cut-off".
- The impact of the energy recovery process shall be added to the inventory results of the product that uses the energy.

### Example 3: Energy recovery with several product systems (cut-off approach)

Organic solvent waste from the manufacturing process of the product A is treated in a waste incineration process with energy recovery on-site and under operational control. The recovered energy is not used in the manufacturing process of product A. It is used in the manufacturing of product B.

Following the cut-off approach, the impact of the waste treatment process shall be allocated to the user of the energy, product B. No impact from the production process for product A shall be allocated to the PCF of product B. If any of the processes, e.g. the production process "Product B" is not operated by company A but operated by a third party, the same approach shall be applied.

Figure 5.9 Energy recovery from waste incineration with application of the cut-off approach



$$\begin{split} & \text{PCF}_{\text{Product A}} = 2.0 \, \text{t CO}_2 \text{e} \, / \, \text{t} \\ & \text{PCF}_{\text{Product B}} = 2.0 \, \text{t CO}_2 \text{e} \, / \, \text{t} + 0.1 \, \text{t CO}_2 \text{e} / \, \text{t} = 2.1 \, \text{t CO}_2 \text{e} / \, \text{t} \\ & \text{PCF}_{\text{Energy}} = 0.1 \, \text{t CO}_2 \text{e} \, / \, 0.2 \, \text{MWh} = 0.5 \, \text{t CO}_2 \text{e} / \, \text{MWh} \end{split}$$

### Following the reverse cut-off approach (waste allocation approach)

- The impact of preparatory steps and supporting activities such as collection, transportation, sorting, dismantling, or shredding shall be added to the inventory results of the product system generating the waste.
- The impact of the process treating waste with energy recovery (e.g. incineration) shall be added to the inventory results of the product system that generated the waste treated in the process.
- The energy recovered from the waste-to-energy process shall be treated as free of burdens. Burdens or credits associated with previous or subsequent life cycles are not considered i.e., are "cut-off".

### Example 4: Energy recovery with several product systems (reverse cut-off approach)

Organic solvent waste from the manufacturing process of the product A is processed by a third party in an energy recovery process. The recovered energy is not used in the manufacturing process of product A. It is used in the manufacturing of product B.

Following the reverse cut-off approach, the impact of the waste incineration process shall be allocated to the generator of the waste, product A. The energy shall be considered free of burden.



Figure 5.10 Energy recovery from waste incineration with application of the reverse cut-off approach



 $PCF_{ProductA} = 2.0 t CO_{2}e/t + 0.1 t CO_{2}e/t = 2.1 t CO_{2}e/t$  $PCF_{productA} = 2.0 t CO_{2}e/t$ 

 $PCF_{Product B} = 2.0 t CO_{2}e/t$   $PCF_{Energy} = 0 t CO_{2}e/MWh$ 

### Following the substitution approach:

The substitution approach is a method to distribute the impacts of multifunctional process (e.g. waste treatment with energy recovery) between the waste generating and energy using system. Following the substitution approach this is achieved, with the help of including a reference system for energy production. Following this approach:

- The impact of preparatory steps and supporting activities such as collection, transportation, sorting, dismantling, or shredding shall be added to the inventory results of the product system generating the waste.
- The energy recovered from the recovery process (e.g. incineration) shall get a PCF representing the impact of the reference energy production (e.g. steam from natural gas of a combined heat and power plant).

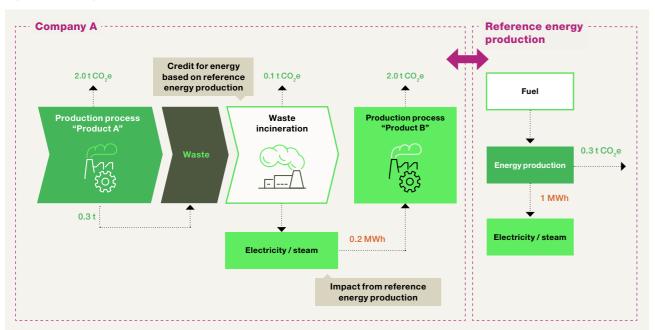
This impact shall be added to the product system using the energy. The product system using the energy receives no benefit from waste treatment with energy recovery.

 The impact of the recovery process (e.g. incineration) shall be added to the waste generating systems. A credit shall be subtracted for the amount of energy recovered using the impact of the reference energy production.

### Example 5: Energy recovery with several product systems (substitution approach)

The production process of product A generates a waste (e.g. solvent waste). This waste is incinerated with energy recovery. The energy is used in the production of product B. As reference, energy can be produced by incineration of a primary fuel.

Figure 5.11 Energy recovery from waste incineration with application of the substitution approach



### Example 6: Energy recovery in a heat network (comparison of the three approaches)

For a comparison of the different approaches, this example is calculated for all three approaches discussed in this chapter. The example shows a simplified scheme of a possible production network in a value chain. The different PCF values for steam and the products calculated with the different approaches are shown in Table 5.5.

Company A produces product A. Waste that is generated in the production of product A is incinerated with energy recovery. In addition to steam generated by the waste incineration with energy recovery, the steam grid consists of a combined heat and power plant and a municipal waste incineration that incinerates product C at its end of life with energy recovery. Both company A and B are using steam in the production of their products. 1t of product A and 1t of Product B are produced in the system. 1t of product C is treated at its end of life.

Figure 5.12 Example of interlinked system with energy recovery from both production and municipal waste

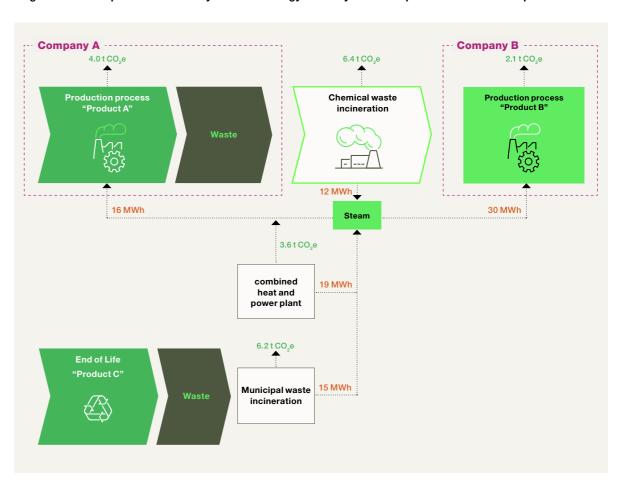




Table 5.5 PCF calculation for example of Figure 6 for the different assessment approaches

Unity t CO o	/ka (matarials)	Cut-off	Reverse cut-off	Substitution
	/kg (materials) /MWh (steam)		approach	
t CO <sub>2</sub> e	/ Ivi vvii (Steaili)	approach	арргоасп	approach
Steam	PCF (Steam, combined heat and power plant)	3.6 / 19 = 0.19	3.6 / 19 = 0.19	3.6 / 19 = 0.19
	PCF (Steam, chemical waste incineration)	6.4 / 12 = 0.53	0	0.19 = PCF (Steam, combined heat and power plant)
	PCF (Steam, municipal waste incineration)	6.2 / 15 = 0.41	0	0.19 = PCF (Steam, combined heat and power plant)
	PCF (Steam, total)	(3.6 + 6.2 + 6.4) / (19 + 15 + 12) = <b>0.35</b>	3.6 / (19 + 15 + 12) = <b>0.078</b>	0.19 = PCF (Steam, combined heat and power plant)
Product A	Direct process emissions	4.0	4.0	4.0
	Waste incineration emissions	0	6.40	6.40
	Steam emissions	16 * 0.35 = 5.63	16 * 0.078 = 1.25	16 * 0.19 = 3.04
	Steam credit	0	0	12 * 0.19 = 2.28
	PCF (Product A)	9.63	11.65	11.16
Product B	Direct process emissions	2.10	2.10	2.10
	Waste incineration emissions	0	0	0
	Steam emissions	30 * 0.35 = 10.56	30 * 0.078 = 2.34	30 * 0.19 = 5.70
	PCF (Product B)	12.66	4.44	7.80
Product C	EoL emissions	0	6.20	6.2 – 15 * 0.19 = <b>3.35</b>

### Guidance for calculating emission factors for material recycling

Material recycling processes are processes that derive a secondary material from a waste material which is further used as material for manufacturing of products. Such processes include chemical recycling through pyrolization, distillation of materials or mechanical recycling. The impact of material recycling shall be included in the product life cycle inventory and system boundary following the calculation approach outlined in this sub-chapter.

### Recycling within the system boundaries of a product

If all processes related to recycling from waste are included in the system boundary, no specific considerations are required. The impact of the recycling process shall be included in the PCF. This approach is described for waste treatment with energy recovery in example 2.

### Recycling outside the system boundaries of a product

Industrial materials can also be recycled along a value chain. Waste material is part of the life cycle of a product system and is reused or recycled as a secondary material in a new product system. This creates the need to split the impact of the processes related to recycling, as they may be shared between two different product life cycles.

To reduce the emission of GHGs, the chemical industry should strive to keep carbon in a material loop. This is primarily achieved through the reduction of waste generation and material recycling of remaining waste. The impact allocation approach should be designed to incentive both.

The different available calculation approaches have been discussed among TfS group members and no consensus has been reached so far. The discussion to select the most appropriate guidance in this chapter will be continued, inviting additional stakeholders to contribute. The guideline will be updated accordingly in due time to reflect changes and consensus. TfS also encourages the development of targeted solutions for such cases through among others, product category rules.

Standards for Product LCAs and corporate sustainability reporting are currently not harmonized and do not fully address the steering effect of PCFs for important technologies with the potential to de-fossilize the chemical industry, such as chemical recycling. The following methodologies are a proposal by the chemical industry to steer those technologies but are not yet harmonized with existing standards, including the GHG Protocol.

Energy intensive recycling (e.g., chemical recycling) technologies are used to recycle waste streams which cannot be recycled through other methods (e.g. mechanical recycling due to technical and economic reasons). Examples are various types of mixed plastics waste after the sorting step and separating materials that cannot be handled in e.g., mechanical recycling. If a recycling technology enables waste to be used as a feedstock (and thus prevents other less favorable end-of-life options and keeps carbon in the loop), it creates societal benefits in form of CO2 reduction and resource savings and should be acknowledged accordingly.

#### The following general rules shall apply:

- Whenever applicable and possible, process subdivision shall be used to divide common processes to avoid the need for allocation. [GHG Protocol Product Life Cycle accounting standard].
- For secondary material derived from a recycling process, whenever available, "allocation methods in line with published and accepted product category rules (PCR) of analogous processes shall be applied, e.g., Plastics Europe" [Pathfinder Framework (PACT powered by WBCSD)].
- 3. If none of the above apply, the two calculation approaches described below shall be consulted.

The first choice shall be a cut off approach due to the requirements of the GHG Protocol [[GHG Protocol Product Standard] with additional requirements on reporting. When providing a cradle-to-gate PCF, the figure for end-of-life emissions shall be reported additionally.

For specified cases, an upstream system expansion approach can be used as an alternative option. In this approach, the cradle-to-gate PCF is provided considering a credit for the avoided waste treatment from the first life cycle.

Both methods are explained in the following text providing examples.

### Following the cut-off approach (also known as recycled content approach):

- The impact of preparatory steps and supporting activities such as collection, transportation, sorting, dismantling, or shredding shall be added to the inventory results of the product system generating the waste.
- The waste input to the recycling process shall be treated as free of burdens. Burdens or credits associated with material from previous or subsequent life cycles are not considered, i.e., they are "cut-off".
- The impact of the recycling process shall be added to the inventory results of the product that uses the secondary material.
- For the product in scope the PCF of all burden shall be reported. Additionally, the EoL of the virgin alternative should be shown in comparison to the recycled product. This is a specific PCF covering EoL effects as well. With this approach, benefits of the recycling of materials can be shown but are beyond a cradle-to-gate scope.

Details of this calculation approach are shown in example 3 of this chapter.

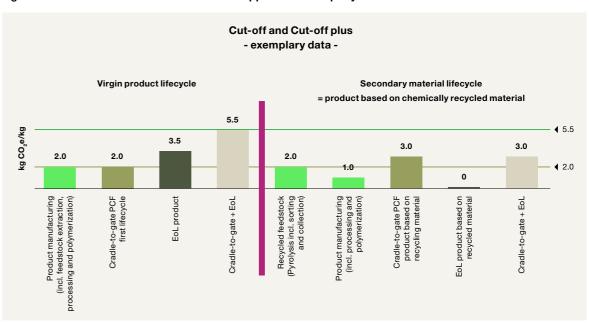
### **Example Cut-off and additional information**

Standard reporting for cut off as follows: PCF virgin (cradle-to-gate first life cycle) =  $2.0 \text{ kg CO}_2\text{e}$  /kg PCF secondary material (cradle-to-gate first life cycle) =  $3.0 \text{ kg CO}_2\text{e}$  /kg

27

Additional reporting information: PCF virgin product incl. EoL =  $5.5~kg~CO_2e/kg$  PCF secondary material incl. EoL =  $3.0~kg~CO_2e/kg$ 

Figure 5.13 Cut-off and additional information approach - exemplary data





The assumed EoL technology in this example for the virgin material was incineration in Europe based on the C-content of the virgin material. All impact of the incineration was allocated to the EoL including the substitution of the recovered energy. If no further information of the EoL of the virgin material is available, the country mix of disposal technologies of the country of origin shall be considered

This approach is close to the cut-off approach described in the GHG Protocol. Through the additional information of the cut-off plus, the benefit of the recycled material compared to a virgin material becomes apparent.

#### Following the Upstream System Expansion (USE) approach:

In exceptional cases the benefits of a recycled material can be shown using the "Upstream System Expansion (USE)" approach [BASF (2020)]. These exceptional cases shall fulfill all the following criteria:

- Showing a societal benefit in form of overall reduced GHG emissions in comparison to relevant other available treatment methods.
- · Being a new technology with high likelihood of improvement of efficiencies after commercial scale up.
- Ensuring the use of regularly updated data according to the TfS guideline.
- Market for the alternative waste treatments is known, the requirements shall be clearly defined.
- ISO compliant substitution approach is applied, the exact use of the waste is known.
- Substitution shall only be applied if the alternative treatment directly replaces the final disposal, and the process is therefore reduced through provision of the co-product.
- Data about the impact of the alternative production process needs to be obtained to calculate the PCF of the alternative product and compare it to the system under study.
- A clear description of the process for selecting the final. EoL option substituted by chemical recycling shall be documented.

The burdens from collection, sorting, recycling step (e.g., pyrolysis) and further processing of the final product (e.g., cracking) are accounted to the secondary material as well the burden of the recycling process. All burdens shall be reported. Additionally, the credit of the displaced EoL impact can be deducted. As a basis for EoL impact estimations, the country mix of disposal technologies of the country of origin shall be considered if there are no further information of the EoL of the virgin material available.

In a second step, the emission of the counterfactual scenario (what would have happened with the waste if not used for recycling) must be identified. In the case of chemical recycling, the used waste streams are difficult to recycle and would have been incinerated otherwise. The emissions of the counterfactual scenario need to be calculated, e.g., incineration of mixed plastics including energy recovery using commonly available technologies in the defined region [GHG Protocol Product Standard (2011)].

The final PCF of the chemically recycled products results from the burdens of the recycling deducted by the savings of the counterfactual scenario, because the technology is benefiting to societal CO<sub>2</sub> savings by replacing the less favorable waste treatment option.

With this approach, benefits of the recycling of materials can be shown but are beyond a cradle-to-gate scope.

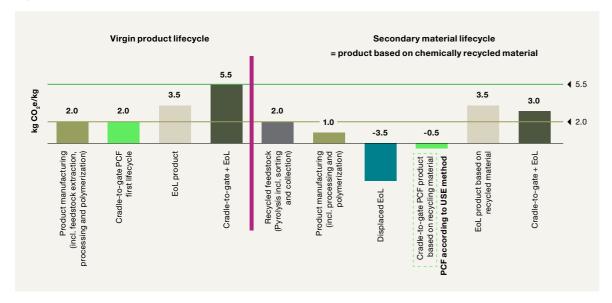
#### Example USE

PCF virgin (cradle-to-gate first life cycle) = 2.0 kg CO<sub>s</sub>e /kg PCF secondary (cradle-to-gate based on recycled mat.) = -0.5 kg CO<sub>2</sub>e /kg

Additional information: PCF virgin product incl. EoL = 5.5 kg CO<sub>2</sub>e /kg

PCF secondary material incl. EoL = 3.0 kg CO<sub>2</sub>e /kg





Depending on the methods used, the corporate accounting in categories 3.1. and 3.12. may differ and are explained in the description of corporate reporting of TfS.

This approach is different to the existing GHG Protocol approach. The results of the USE method incl. EoL considers a scope beyond cradle-to-gate. To derive a PCF from there can be further addressed in a stakeholder alignment process. The accounting for the EoL along the value chain among the recyclers and users of the material should be a part of this.

#### 5.2.8.5 Direct emissions

Direct emissions are emissions from processes owned or controlled by the company arising from:

- · Chemical reactions.
- · Waste treatment with and without energy use (e.g., flares).
- Fuel and residues incineration in process plants.

Direct emissions shall be calculated by determining the amount of emitted GHGs based on stoichiometry, mass balance or measured data. The emissions shall then be multiplied with the respective global warming potential (GWP) to calculate the emission factor as CO<sub>2</sub>eq per declared unit. When relevant, fossil and biogenic direct CO e emissions to be reported separately according to the guidance in chapter 5.2.10.1.

### 5.2.9 Multi-output processes

This chapter is about attributing inputs and emissions in multi-output situations, i.e., when a process delivers more than one product, referred to as co-products. The term co-product also includes energy products such as steam or electricity, or any other product with a defined economic value such as a residual fuel. Energy is in this sense understood as direct energy e.g., from exothermal reactions [Pathfinder Framework (PACT powered by WBCSD)]. Waste materials that go directly to a final disposal, e.g., an incineration or landfill, are not co-products as they do not have an economic value and hence, shall be excluded from the attribution of environmental burdens of the multi-output process. The energy generation from waste incineration is described in the waste treatment chapter.

Leaning on hierarchies described in the [GHG Protocol Product Standard]. ISO 14040 [ISO 14040:2006]. ISO 14044 [ISO 14044: 2006], ISO 14067 [ISO 14067: 2018], [Pathfinder Framework (PACT powered by WBCSD)]. and the European Commission Environmental Footprint recommendations, the following steps shall be applied to attribute impacts in multi-output situations (see Figure 5.15):

1) Multi-output situations shall be avoided by using process subdivision, whenever possible. The common process shall be disaggregated into sub-processes that separately produce the co-products. Process subdivision may be done through sub-metering specific process lines and/or using engineering models to model the process inputs and outputs [GHG Protocol Product Standard].

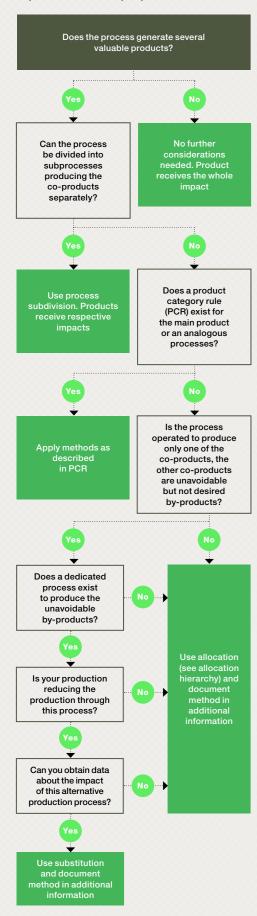
- 2) If the multi-output situation cannot be avoided by subdivision, a system expansion or allocation following the approach described in published and accepted product category rules (PCR) or Industry Association projects, where available, for corresponding product systems shall be applied (see 5.2.4 Standards used). When more than one PCR exists for a product or product category, priority shall be given to allocations rules as described in chapter 5.2.9.3.
- 3) System expansion and substitution can be a means of avoiding allocation. The product system that is substituted by the co-product is integrated in the product system under study. In practice, the co-products are compared to other substitutable products, and the environmental burdens associated with the substituted product(s) are subtracted from the product system under study [ISO 14044: 2006]. System expansion refers to expanding the system by including the co-products into the system boundary and communicate PCF results for the expanded system [PEF - GUIDE: 2012]. Only system expansion by substitution (further referred to as "substitution") is acceptable as the declared unit stays as defined in chapter 5.1.3.

Substitution, as described in chapter 5.2.9.1, may be applied to attribute impact to co-products in multi-output situations if all the following apply:

- a. The co-products are generated in the process additionally but are not the main products of the process. Main products are defined as products that the process is operated for and optimized to produce. Additionally, the economic values of the main products are generally significantly higher than for the co-products.
- b. The co-product directly replaces an alternative product with a dedicated production process on the market. The production of this alternative product is reduced through provision of the co-product.
- c. Data about the impact of the alternative production process is available to calculate the PCF of the alternative product.
- 4) In all other cases companies shall allocate the impact to co-products following the allocation rules described in chapter 5.2.9.3.

The applied approach to solve multi-functionality shall always be stated and justified.

Figure 5.15 Decision tree to attribute impacts in multi-output processes



### 5.2.9.1 Substitution

In Substitution, the co-products of process are compared to similar alternative products, and the environmental burdens associated with the alternative product(s) are subtracted from the product system under study to obtain the impact of the main product of the production process (see Figure 5.15) [ISO 14044: 2006].

The use of substitution as a means to avoid allocation requires an understanding of the market for the co-products. To ensure that an ISO compliant substitution approach is applied, the exact use of the co-product needs to be known. Substitution shall only be applied if the co-product directly replaces the alternative product on the market and the production of this alternative product is therefore reduced through provision of the co-product. Data about the impact of the alternative production process needs to be obtained to calculate the PCF of the alternative product and subtract it from the system under study.

A clear description of the process for selecting the alternative product substituted by the co-product shall be documented.

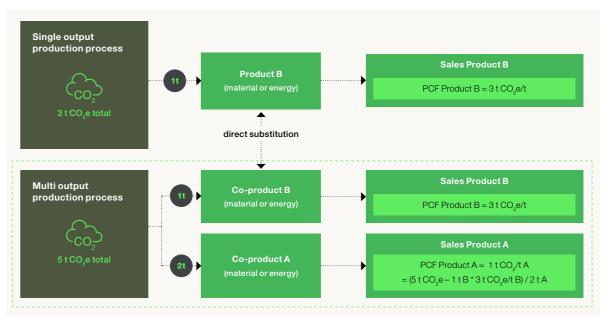
Energy co-products such as residual fuels or excess steam shall be treated by substitution if these co-products substitute products that would have been otherwise generated from a primary energy source. Please see further explanation in below example.

### 5.2.9.2 Examples for Substitution

In the example both co-product A and co-product B are produced as co-products of the same process. The process produces 2 t co-product A and 1 t co-product B with associated CO<sub>2</sub>e emissions of 5 t CO<sub>2</sub>e (see Figure 5.16). Process subdivision is not possible, and a product category rule does not exist. The process is operated and optimized to produce co-product A as the main product. Co-product B is unavoidably co-produced and is considered a by-product. The co-product B is the same product as product B derived from a single output production process and substitutes product B (material or energy) from a single-output process.

In the market, co-product B directly substitutes an alternative product B, produced through a process with an impact of 3 t  $\rm CO_2e/1$  t product B. This impact is now assumed for co-product B from the system under study. As the process under study produces 1 t of product B within the system boundaries, the impact of the substituted alternative process can be subtracted from the total impact of the process. As a result, 2 t of co-product A have an impact of (5-3) t  $\rm CO_2e = 2$  t  $\rm CO_2e$ . As a result, co-product A has a PCF of 1 t  $\rm CO_2e/t$  co-product A.

Figure 5.16 Substitution and its modelling of multi output processes



### 5.2.9.3 Allocation rules

Allocation means splitting multi-output processes into single output unit processes using physical, economic, or other criteria by partitioning the input and output flows of a process or a product system between the product system under study and one or more other product systems. When outputs include both co-products and waste, the inputs and outputs shall be allocated to the co-products only.

There are different allocation methods applicable for the case of a multi-output process. ISO 14067 [ISO 14067: 2018] differentiates between allocation based on the underlying physical relationships between the products and co-products such as mass, volume or energetic content and economic allocation – where physical relationship is the preferred choice. Furthermore, input materials as e.g. chemicals can be allocated by stoichiometry to the products according to the chemical reaction and elemental connectivity.

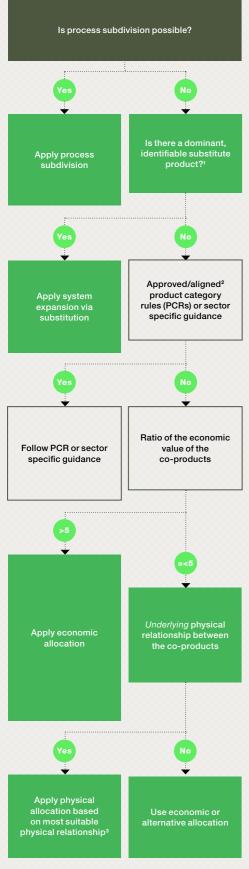
The following general rules shall apply:

If the multi-output situation cannot be avoided, emissions shall be divided among the co-products in an accurate and consistent manner. This is essential for the quality of a PCF. Allocation rules shall follow the hierarchy described below [Pathfinder Framework (PACT powered by WBCSD)]:

- a) Allocation methods in line with published and accepted product category rules (PCR) of analogous processes shall be applied where available (see 5.2.4 Standards used). When more than one PCR exists for a product or product category, priority shall be given to allocations rules in.
- a. Existing regional law.
- b. PCRs from world-wide operating associations.
- c. PCRs from regionally operating associations. (e.g., Plastics Europe).
- d. PCR from EPD programs.



Figure 5.17 Decision-making tree to show allocation rules and reduce assessment burden downstream [Pathfinder Framework (PACT powered by WBCSD)]



b) The guidance of the WBCSD Chemicals [WBCSD Chemicals LCA Guidance (2014)] used the application of the economic value of co-products as a criterion to decide between physical allocation and economic allocation firstly. The criterion for economic allocation was adopted as well by the Pathfinder project and aligned with TfS (Figure 5.17). Economic allocation factors should be calculated based on stable market prices, as a yearly average or over multiple years in case of high fluctuation (e.g. >100%) of prices to average out high fluctuations of prices, influencing the outcome of an allocation process based on economic values as prices [BASF SE (2021)]. If market prices are not available, other economic factors can be applied.

If the share of a co-product is very small (in mass or volume <= 1%), it can be skipped in the decision about the allocation method (see also Chapter 5.2.3 for rules on cut-off criteria). If there are more than two co-products, use the highest and lowest value of all co-products to determine the value ratio.

Exceptions to the above allocation rules are possible only in rare instances such as:

- Carbon dioxide that is captured and used as input in another process is be treated separately (see chapter 5.2.10.3 Carbon Capture & Utilization).
- 2. If hydrogen is a co-product allocation by heating value shall be applied because of the low molecular weight of hydrogen. Example: Syngas process that generates CO and hydrogen, both are gases and valuable products. If hydrogen is a co-product in a multi-output process, mass allocation shall not be applied because of the low molecular weight of hydrogen.

The applied approach to solve multi-output situations shall always be stated and justified, and the sum of the allocated inputs and outputs of a unit process shall be equal to the inputs and outputs of the unit process before allocation.

The allocation procedure has a significant impact on the PCF result as can be seen below in the example of Chlor-Alkali Electrolysis, a multi-output process generating

5.2.9.4 Examples for allocation

chlorine, caustic soda, and hydrogen (see Figure 5.18).

Hence a uniform approach for how to deal with multi-output situations for all possible types of product and co-products is needed to generate consistent and comparable results.

Figure 5.18 Outputs of a Chlor-Alkali electrolysis process



It should be noted that a PCR exists for the Chlor-Alkali Electrolysis and the different allocation approaches shown are simply illustrative examples.

#### Mass-based allocation

This type of allocation is the distribution according to mass, measured in terms of total mass (see Table 5.6).

Table 5.6 Example calculation for mass-based allocation

Definition	Mass [kg/kg Chlorine]	Share of impact
Chlorine	1.00	47%
Caustic soda (100%)	1.085	51%
Hydrogen	0.028	2%
Sum		100%

<sup>(1)</sup> System expansion via substitution should only be used if there is a dominant, identifiable displaced product and production path for the displaced product based on sector consensus.

<sup>(2)</sup> Sector specific guidance or PCRs shall be used if approved and required as Drop-in standards by TfS for Chemical Industry, by Catena-X for other automotive industry supplying sectors or by WBCSD pathfinder for sectors other than those covered by TfS and Catena-X.

<sup>(3)</sup> In doubt, mass allocation should be prioritized, but there are instances where other allocation factors may be more suitable (e.g. volume for gases, energy content for energy).



### Stoichiometric or elemental allocation

Stoichiometric ratios of chemical reactions can be used as basis for the allocation. This approach is helpful if mass flows would not reflect the elemental reality of the co-products. This allocation can be used

for input materials that have a chemical connectivity only to certain products and not all co-products. Stoichiometric or elemental allocation can be combined with e.g., mass allocation for other raw materials, energy, waste, emissions etc (see Table 5.7).

Table 5.7 Example calculation for stoichiometric or elemental allocation

Definition	Molar mass [g/mol]	Stoichiometric relation to NaCl	Share of NaCl impact	
Chlorine, Cl <sub>2</sub>	70.9	0.5	60.7%	
Caustic soda, NaOH (100%)	40	1	39.3%	
Hydrogen, H <sub>2</sub>	2	0	0%	
Sum 100%				
Share of NaCl impact = Molar mass of product * stoichiometric factor of product / molar mass of NaCl.				

### **Economic allocation**

The economic allocation refers to the economic value of the products at the location (e.g., in the plant) as well as in the state (e.g., not cleaned) and quantity as provided by the multi-functional process. A specific market price is attributed to each product (see Table 5.8). If large fluctuations in prices exist, an average price over several years should be calculated to reduce these fluctuations. Most recent prices should be used if available and appropriate.

Table 5.8 Example calculation for economic allocation

Definition	Value [USD/kg]	Mass [kg/kg Chlorine]	Value x Mass [USD]	Share of impact
Chlorine	0.42	1.00	0.42	63%
Caustic soda (100%)	0.10	1.085	0.1085	16%
Hydrogen	5.00	0.028	0.14	21%
Sum			0.6685	100%

In cases where the product is not sold or the determination of market prices is hardly possible (e.g. intermediates which are internally used, chlorine for PVC), other approaches might be used, e.g. a combination of production costs and market price of the processed product or the turnover.

Overview of calculation examples with a multi-output-allocation

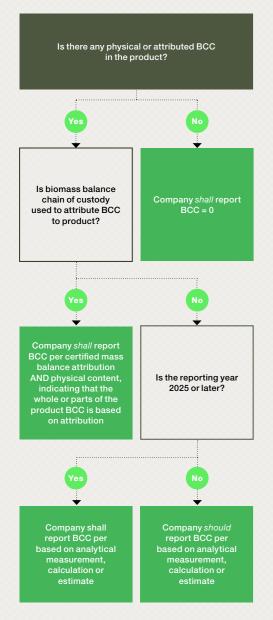
Table 5.9 General examples for allocation approaches and calculation rules

	Applicable PCF calculation rule		
Example case	"how to do it"		
Chlorine-Alkali-Electrolysis delivers besides chlorine, mainly hydrogen and	Follow decision tree above: apply allocation scheme as specified in the PCR from [Eurochlor [2022]].		
sodium hydroxide; energy co-products such as steam are not generated.	Sodium chloride input is allocated by means of stoichiometry to all products containing either sodium or chlorine atoms (or both): chlorine, sodium hydroxide, sodium hypochlorite and sodium sulphate.		
	Sulphuric acid input is allocated to chlorine production only, since it is used for chlorine drying.		
	Hydrogen emissions are allocated to hydrogen production only, since they refer to losses of hydrogen to the atmosphere.		
	Chlorine gas emissions are allocated to chlorine production only, since they refer to losses of chlorine to the atmosphere.		
	Electricity, steam and all other inputs and outputs are allocated by mas to all valuable products, for solutions to mass content of active molecul		
The steam cracker process turns fossil hydrocarbon feedstocks (predominantly ethane, LPG, naphtha, or gas oil) into several different main products, like ethylene and propylene, benzene, butadiene and hydrogen. The process yields additional further chemicals like, acetylene, butene, toluene and xylene.	This complicated process for a LCA needs some specific approach for an accurate calculation. Therefore, a PCR from Plastics Europe¹ was developed to harmonize the approach. The PCR distinguishes per definition between so-called "main products" (ethylene, propylene, benzene, butadiene, hydrogen, toluene, Xylene and butenes) and "additional products" (all other products). It is defined that the feedstock used shall be allocated on mass basis to all stean cracker products.  Energy demand and emissions shall be exclusively allocated on a		
	mass basis to the "main products" only.		
The production of formaldehyde from methanol produces besides formaldehyde excess steam that is used in another production plant at the same site of the reporting company. The steam substitutes steam generated in an on-site CHP plant based on natural gas.	The formaldehyde process produces a co-product which is only use in energy recovery. Following the decision tree and its exceptions, the allocation issue can be solved by system expansion and substitution. This means that the CO2 impact of the inputs and outputs of the process are completely allocated to the main product. At the same time, however, the process receives a CO2 credits that correspond to the CO2 impact of steam generated in the on-site CHP plant based on natural gas. When using the waste steam as input in another production process is carries the CO2 burden of the steam generated in the CHP based on natural gas. In this way the CO2 balance is closed, and the steam generating process is rewarded as it produces a product that substitutes a product that would have been otherwise produced.		
Atmospheric gases as nitrogen, oxygen, argon and other inert gases are produced using a process known as air separation. In this process, atmospheric air is split into its primary components via a fractional distillation. Cryogenic air separation units (ASUs) are built to provide nitrogen or oxygen and often co-produce argon. High purity gases can be obtained from this process. Rare gases as neon, krypton, and xenon can be isolated with the distillation of air using at least two distillation columns.	Follow decision tree above: no PCR exists, comparison of economic values of co-products (=prices) results in a ratio of > 5. [Price Product 1 (max) / price Product 2 (min) > 5]. The CO <sub>2</sub> e impact from the input and output flows shall be allocated based on an economic allocation approach.  If the economic values of co-products (=prices) results in a ratio of =< 5, allocations based on physical relations shall be applied. In a typical distillation process that is applied for the separation of e.g. different chemicals with different boiling points, the boiling points can be used as basis for allocation. Higher boiling points get higher burdens because more energy is needed to distill the products.		
This type of distillation can be transferred to almost all other distillations very often used in the chemical industry. The process is applied for the separation of different fractions of chemicals and for the purification of chemicals.			

 $<sup>(1) \</sup> Plastics \ Europe \ recommendation \ on \ Steam \ Cracker \ allocation. \ Plastic \ Europe \ - \ Stream \ Cracker \ Allocation$ 



Figure 5.19 Decision Tree for Reporting of Biogenic Carbon Content (BCC) in a Product<sup>1</sup>



### Other requirements:

Company shall indicate if BCC is based on physical basis or attribution.

BCC shall be corrected after any economic allocations applied in supply chain.

(1) 2025 was set as the first mandatory year for reporting biogenic carbon to give all involved companies enough time to prepare for this.

### 5.2.10 Additional rules and requirements

### 5.2.10.1 Approach to consider biogenic carbon in the PCF

"During photosynthesis, plants remove carbon (as CO<sub>2</sub>) from the atmosphere and store it in plant tissue. Until this carbon is cycled back into the atmosphere, it resides in the carbon pools" [GHG Protocol Corporate Standard], like bio-based materials. "Carbon can remain in some of these pools for long periods of time, sometimes for centuries. An increase in the stock of sequestered carbon stored in these pools represents a net removal of carbon from the atmosphere". As biobased materials origin from plants, the same is true for them."

The requirements in this guidance are aligned to the requirements set out in ISO 14067 [ISO 14067: 2018].

According to ISO 14067, **biogenic removals from** CO<sub>2</sub> **uptake** during biomass growth shall be included in the PCF calculation. Additionally, all biogenic emissions (e.g. methane emissions from manure application etc.) and further emissions from relevant processes, such as cultivation, production and harvesting of biomass shall be included in the PCF [ISO 14067: 2018].

Removals of CO<sub>2</sub> into biomass shall be characterized in the PCF calculation as –1 kg CO<sub>2</sub>/kg CO<sub>2</sub> when entering the product system, while biogenic CO<sub>2</sub> emissions shall be characterized as +1 kg CO<sub>2</sub>e/kg CO<sub>2</sub> of biogenic carbon [ISO 14067: 2018]. As referred to in Chapter 5.3.2, the PCF, that considers biogenic emissions and removals shall be reported as **PCF** (including biogenic CO<sub>2</sub> removal).

It should be noted that other systems (namely the European Commission Product Environmental Footprint (PEF 2021) system) treat biogenic emissions and removals differently. PEF does not consider biogenic CO<sub>o</sub> emissions and biogenic CO<sub>2</sub> removals (0/0 approach) so far, but biogenic CH, emissions, Furthermore, PEF and the IGHG Protocol Product Standard consider biogenic CO, emissions and biogenic CO<sub>2</sub> removals as neutral, independently from its end-of-life treatment. For short term uses of materials with incineration, this approach is identical with the approach of consideration of biogenic carbon uptake and subsequent emission from incineration. To fulfill PEF and current GHG Protocol requirements, additionally the "PCF (excluding biogenic CO<sub>2</sub> removal)" shall be reported, which does not consider biogenic removals, but all biogenic and fossil emissions. The biogenic emissions contain the CH, emissions that are derived from bio-based C and converted to Methane as well and are transferred to CO.e. N.O emissions derived from bio-based materials are expressed in CO<sub>a</sub>e as well. If N<sub>a</sub>O is emitted from the use of a fertilizer that is based on fossil materials it is linked to the fossil CO<sub>o</sub>e.

The upcoming GHG P Land sector and removal Guidance will overrule all the existing GHG P standards in terms of biogenic emissions and accounting requirements. TfS will update this guideline if the final version is published.

Because the prescribed scope of PCF (including biogenic uptake) within this context guideline is a cradle-to-gate consideration exclusively, the total carbon content and the biogenic carbon content of the material shall also be reported alongside the PCF (including biogenic uptake) with the aim to close the biogenic carbon balance in further downstream calculations or at the end-of-life, which are not in scope of this document [BASF SE (BASF)],

[ISO 14067: 2018]. Figure 5.19 presents a decision tree for biogenic carbon content (BCC) reporting. Biogenic carbon is defined as carbon derived from biomass. Biomass refers to material of biological origin and includes both living and dead organic material, such as trees, crops, grasses, tree litter, algae, animals, manure, and waste of biological origin. In this document, peat is excluded from the definition of biomass [ISO 14067: 2018]. Within the context of products, biomass-derived carbon contained in a product is referred to as the biogenic carbon content of the product [ISO 14067: 2018]. BCC may be present in products due to physical presence or due to attribution in biomass balance. If biomass balance is used then provisions shall be in place to avoid double counting, especially in products which do not receive attributed BCC.

If the mass of biogenic carbon containing materials in the product is less than 5% of the mass of the product, the declaration of biogenic carbon content may be omitted ([EN15804+A2 2019: 46]).

An example of how to calculate and report the biogenic uptake and the carbon content is presented for a bio-based ethanol below.

- Carbon content in ethanol (carbon molecular weight / total ethanol molecular weight) = (24g/mol / 46g/mol) = 52.17% C content in ethanol.
- 1 kg ethanol contains 521.7 g C.
- As the biogenic Carbon content accounts 100%, the biogenic C content is also 521.7 g C/kg.
- The biogenic removal is 521.7 g C/kg \* 44/12 (conversion of carbon into carbon dioxide)
   = 1 913 g CO<sub>2</sub> / kg ethanol.

When the ethanol is incinerated e.g. in an EoL process, this amount of CO<sub>2</sub>e will be released as emission¹. If the ethanol is used as a pre-cursor for a chemical product and this product is applied in a long-term application, the contribution from the ethanol is negative. The new GHG Protocol Land sector and removal Guidance has a new approach on how to account for delayed emissions from product carbon pools. The TfS guideline will be adapted if the Guidance is published.

An example how to report emissions for biobased ethanol is provided below in table 5.10.

Table 5.10 Calculation and reporting of PCF results with biogenic materials included

Simplified calculation example: For 1 kg of ethanol	According to ISO 14067: 2018; GHG Protocol Product Standard	According to PEF 2021
Biogenic carbon in products (kg biogenic C/kg ethanol)	0.521	0.521
Equivalent biogenic carbon removal in product, expressed in carbon dioxide (kg CO <sub>2</sub> /kg ethanol)	-1.9	0.0
Equivalent biogenic carbon overall removal, expressed in carbon dioxide (kg CO <sub>2</sub> /kg ethanol)	-2.334	0.0
Emissions, land use and direct land use change (kg CO <sub>2</sub> e/kg ethanol)	0.2	0.2
Of that is direct land use change (kg CO <sub>2</sub> e/kg ethanol)	0.1	0.1
Emissions, biogenic (kg CO <sub>2</sub> e/kg ethanol)	0.8 (0.4 from Methane)	0.4 (Methane)
Emissions, fossil (kg CO <sub>2</sub> e/kg ethanol) (net result of fossil emissions and fossil removals)	2.0	2.0
Cradle-to-gate emission (kg CO <sub>2</sub> e per kg ethanol	-2.334+0.2+0.8+2.0 = <b>0.67</b>	0.0+0.2+0.4+2.0 = <b>2.6</b>

<sup>(1)</sup> During modeling of EoL, e.g. when biomass is used as energy source for a process, the biogenic carbon in the product should be released in the same way like the fossil carbon depending on the specific EoL technology (e.g. under consideration of conversion into all relevant carbon-based gases (CO<sub>2</sub>, CO, CH<sub>2</sub>)). It should be checked that the carbon balance i closed (uptake equal emissions).



- Biogenic methane emission and corresponding CO<sub>2</sub> uptake:
- 0.4 kg CO<sub>2</sub>e / kg ethanol
- 0.4 / GWP factor methane (30 kg / kg Methane) = 0.013 kg Methane / kg Ethanol
- 0.013 kg methane = (0.013/16) \* 44 = 0.04 kg CO<sub>2</sub> uptake
- Additional uptake from biogenic CO<sub>2</sub> emission: 0.4 kg CO<sub>2</sub>e / kg ethanol
- Total CO<sub>2</sub> uptake:
- $-1.9 \text{ kg CO}_2 0.04 \text{ kg CO}_2 0.4 \text{ kg CO}_2 = -2.34 \text{ kg CO}_2$

According to ISO 14067 [ISO 14067: 2018] the biogenic carbon in products, fossil and biogenic GHG emissions and removals shall be reported. GHG emissions and removals from land use should be reported.

In some cases, e.g. when allocation is applied, the carbon flows might not represent physical reality in terms of C-content. To avoid misleading or incorrect calculations, a carbon correction shall be applied at the end of the PCF calculations. It must be ensured that the biogenic carbon content in the product is equal to the sum of biogenic removal of  $\mathrm{CO}_2$  and biogenic emissions of  $\mathrm{CO}_2$  and methane. If this is not the case (e.g. because of allocation somewhere along the value chain) then the value of the biogenic  $\mathrm{CO}_2$  removal shall be adjusted.

Consequently, the information shown in Table 5.10 needs to be reported and transferred to the recipient separately (see also Chapter 5.3). In addition, information about carbon content shall be added:

- Biogenic carbon content: 0.5217 kg C / kg Ethanol.
- Total carbon content: 0.5217 kg C / kg Ethanol
   (= biogenic carbon content of 0.5217 kg C / kg product + fossil carbon content of 0 kg C / kg product).

For the raw material calculation in section 5.2.8.2 the total figures according to ISO 14067 [ISO 14067: 2018] shall be used. The results for a product calculation includes the biogenic removal at the gate. The biogenic carbon uptake shall be reported in addition. This will enable the calculation of a correct PCF depending on the end-of life treatment of the final user of the product.

When considering biogenic carbon removal in products for a specified duration, the effect of the timing of GHG emissions and removals shall be assessed [ISO 14067: 2018].

Where GHG emissions and removals arising from the use stage and/or from the end-of-life stage occur over more than 10 years (if not otherwise specified in the relevant PCR) after the product has been brought into use, the timing of GHG emissions and removals relative to the year of production of the product shall be specified in the life cycle inventory. The effect of timing of the GHG emissions and removals from the product system (as  $\mathrm{CO}_2\mathrm{e}$ ), if calculated, shall be documented separately in the inventory [ISO 14067: 2018].

The biogenic carbon content of the packaging (if considered in the PCF) shall be excluded or reported separately for an accurate end-of life calculation.

Biomass used for chemical production should be of high quality and should be produced addressing important sustainability aspects of a high level of sustainability. The following requirements should apply for the usage of mass balance chain of custody in determination of PCF:

- The biomass used should follow a transparent certification standard and the conformance to the certification should be verified by an independent and qualified independent party.
- a. The certification system shall have clear chain of custody rules, traceability requirements, defined boundaries, guidelines for marketing claims, include safeguards against double counting in any sense, and shall identify the type of sustainable raw material throughout the supply chain.
- Examples of acceptable certification systems include ISCC PLUS, REDcert2, UL ECVP 2809, RSB Advanced Materials, FSC, RSPO, or equivalent.
- The LCA of the manufacturing process in which the mass balance attribution occurs can be reviewed by an independent party and confirmed to be in conformance with ISO 14044 [ISO 14044: 2006] or ISO 14067 [ISO 14067: 2018]. The study shall document how the material flow and attributions were calculated.

For example, the EU sustainability criteria are extended to cover biomass for heating and cooling and power generation in the revised Directive [EU] 2018/2001. EU countries were required to transpose the new rules by 30 June 2021, and the voluntary schemes have to adjust the certification approaches to meet the new requirements.

For a scheme to be recognized by the European Commission, it must fulfil criteria such as:

- Feedstock producers comply with the sustainability criteria of the revised Renewable Energy Directive and its implementing legislation.
- Information on the sustainability characteristics can be traced to the origin of the feedstock.
- · All information is well documented.
- Companies are audited before they start to participate in the scheme and retroactive audits take place regularly.
- The auditors have both the generic and specific auditing skills needed with regards to the scheme's criteria.
- The decision recognizing a voluntary scheme has usually a legal period of validity of 5 years.

If a mixed raw material containing less than 100% biogenic materials is used, the biogenic content shall be calculated according to the share of the biobased materials and reported accordingly. The other share of materials is linked to fossil Carbon.

If a PEF compliant calculation is requested, the PEF figures shall be used.

### 5.2.10.2 Land-use-change emissions

Land use change (LUC) refers to a change from one land use (can be natural habitats such as primary forests or also agricultural land) to another land use (most times to "human use or management of land."). As a result of land use change, GHG emissions and removals occur

through changes in soil- and above- and below groundbiomass carbon stocks that are not the result of changes to management of land [ISO 14067: 2018]. Changes in management of land within the same land-use category are not considered land use change (e.g. agricultural land to agricultural land). Land use change can be classified as direct or indirect land use change (Table 5.11):

### Table 5.11 dLUC and iLUC [ISO 14067: 2018]

### Direct land use change (dLUC)

Change in the human use of land within the relevant boundary which leads to a change in soil- and biomass carbon stocks.

E.g. Primary forest is converted to agricultural land or grassland.

GHG emissions and removals associated with these changes from reference land use to land use under assessment need to be addressed and shall be included in the PCF calculation.

### Indirect land use change (iLUC); optional

Change in the use of land, which is a consequence of direct land use change, but which occurs outside the relevant boundary.

E.g. Change in use of agricultural land for food to agricultural land for bio-based chemical feedstocks which lead to shift of food production outside the boundary.

In accordance with ISO 14067 [ISO 14067: 2018] GHG emissions and removals occurring because of dLUC shall be included in the PCF calculation and shall be declared separately in the documentation [ISO 14067: 2018]. GHG emissions and removals as a result iLUC can be considered for inclusion and – if calculated - shall be documented separately [ISO 14067: 2018].

The **GHG** emissions and removals occurring because of dLUC within the **last decades** (IPCC tier 1 period of 20 years is frequently used) shall be assessed in accordance with internationally recognized methods, such as **the IPCC Guidelines for National Greenhouse Gas Inventories** [IPCC- GHG Inventories Guidelines].

If a specific approach (e.g. based on site, regional or national data) is used, the data shall be based on a verified study, a peer reviewed study or similar **scientific evidence** and shall be documented in the PCF study report [ISO 14067: 2018].

If a product is 100% fossil based including all relevant pre-cursors, this category is of very low relevance and can be neglected in the evaluation and should be reported as "not applicable".

### 5.2.10.3 Avoided emissions and offsets

### **Definition of avoided emissions**

In this standard, avoided emissions are quantified as emissions reductions that are indirectly caused by the studied product or process or by market responses to the studied product or process that occurs in the studied product's life cycle. Avoided emissions shall not be subtracted from the total inventory results of the PCF.

For more information on avoided emissions see WRI Guideline on avoided emissions [Estimating and Reporting the Comparative Emissions Impacts of Products], [GHG Protocol Product Standard], [ICCA - Avoided Emission Challenge [2017]] or [WBCSD - SOS 1.5], expected to be released end of 2022.

### Definition of emission offsets

"Emission offsets are emission credits (in the form of emission trading or funding of emission-reduction projects) that a company purchases to offset the impact of the studied product's emissions. Companies typically use offsets for one of two reasons: to meet a reduction goal that they cannot reach with reductions alone, or to claim a product as carbon neutral" [GHG Protocol Product Standard].

Emission offsets shall not be subtracted from the total inventory results of the PCF. However, if a company wishes to purchase offsets for its product inventory, it may provide information on the offsets separately from the inventory results. For these offsets to be provided separately as additional information, the company should: Purchase offsets for which GHG emission benefits are quantified following internationally accepted GHG mitigation project accounting methodologies (e.g. GHG Protocol Project Protocol); only account for product-level offsets to avoid double counting of corporate-level offsets [GHG Protocol Product Standard].

### Definition of emission removals

The sequestration or absorption of GHG emissions from the atmosphere, which most typically occurs when  $\mathrm{CO}_2$  is absorbed by biogenic materials during photosynthesis.



Table 5.12 Examples for avoided emissions by off-setting

Example case	Applicable PCF calculation rule	Voluntary additional information for emission offset
The company purchases emission credits from a project investing in reforestation to offset 50% of the calculated PCF	The PCF remains the same as calculated	The emissions offset of 50% may be provided separately from inventory results
The company purchases emission credits from a carbon capture and storage facility from another company to offset 30% of the calculated PCF	The PCF remains the same as calculated. The GHG reduction by CCS cannot be considered as emission reduction in the PCF, as the CCS is not part of the product system	The emissions offset of 30% may be provided separately from the inventory results

Since there are developments towards new ISO standards, aspects might be addressed differently. On ISO Level there is a new standard, ISO 14068 "Climate neutrality" under development. A Net-Zero approach of ISO started as well with the IWAR 42 Net Zero Guiding principles. These activities might initiate new calculation aspects and implementation of PCF in specific calculations. This guideline will be updated accordingly, when these standards will be published, and new requirements need to be addressed.

### 5.2.10.4 Carbon Capture followed by Storage, Utilization

"Carbon Capture" refers to processes where  $\mathrm{CO}_2$  is separated from industrial and energy-related sources or technically captured from the atmosphere. This guidance refers to the capturing of  $\mathrm{CO}_2$  at the emissions source only. Direct air capture technologies are out of the scope of this subchapter. For other technologies to capture other Carbon sources (e.g.  $\mathrm{CH}_2$ ), further definitions are needed.

CCS (Carbon Capture and Storage, or more accurate:  $\rm CO_2$  capture and storage) refers to the separation of  $\rm CO_2$ , and injection into a geological formation, resulting in long-term isolation from the atmosphere.

Long-term means the minimum period necessary to be considered an effective and environmentally safe climate change mitigation option [ ISO 27917:2017], [ISO Guide 84:2020].

CCU (CO<sub>2</sub> capture and utilization) refers to technical processes where the separated CO<sub>2</sub> is converted into valuable products. In contrast to CCS, the CO<sub>2</sub> storage in CCU is only temporary. Emissions can be delayed and thus, do not contribute to climate change during the time of storage [Müller, Kätelhön et al (2020)].

CC only refers to industrial emission sources, while biological processes, where  ${\rm CO}_2$  is also stored (or sequestered) such as planting trees, is not covered by the terminology.

### Carbon Capture and Storage

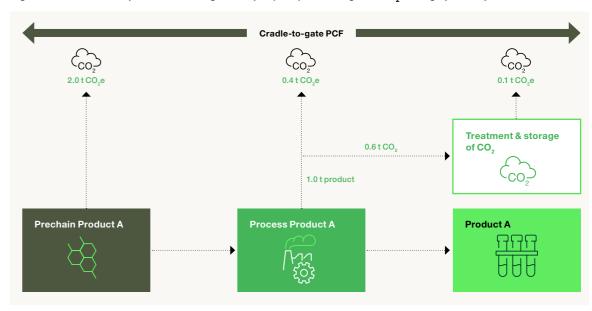
CCS may be included in the PCF calculation if a permanent and complete storage in storage facilities is guaranteed. Permanent storage technologies are characterized by a very low risk of a physical reversal of the storage process. The World Economic Forum offers a comprehensive overview of storing technologies. The net result of GHG emissions, stored GHG emissions and the deployed storage technology shall be documented. The individual amounts of emitted GHG (e.g. via capturing, transport, storage) and stored GHG could be reported separately [BASF SE (2021)].

CSS may only be included in the PCF if the CCS technology is active whenever the product is being produced.

Table 5.13 Examples for CCS

Example case (See figure 5.20)	Applicable PCF calculation rule	Voluntary additional information for emission offset
The company installs a facility for carbon capture and guarantees permanent and complete storage of 0.6 tons of CO <sub>2</sub> (CCS)	The capture of 0.6 tons of CO <sub>2</sub> should be considered. The net result of the PCF shall include the stored emission of 0.6 tons as well as released emissions from the capturing, any transport as well as the storage (See figure 5.20)	Absolute values of released emissions and stored emissions can be reported individually

Figure 5.20 Carbon Capture and Storage example (CCS) assuming 0.6t CO, storage per ton product A



PCF (Product A) =  $2.0 \text{ CO}_2\text{e}/\text{t} + 0.4 \text{ CO}_2\text{e}/\text{t} + 0.1 = 2.5 \text{ t}$ 

Without Carbon Capture and Storage, the emission of the "Process Product A" would be 1t CO<sub>2</sub>e, which would result in an overall emission of 3.0t CO<sub>2</sub>e. With CCS, the emission of the "Process Product A" is lowered to 0.4t. For treatment and storage, 0.1t CO<sub>2</sub>e are emitted; thus, the overall net CO<sub>2</sub>e is 2.5t CO<sub>3</sub>e (2.0t + 0.4t +0.1t)

- Net PCF including CCS (Product A) to be reported: 2.5 t CO.e.
- Voluntary additional information on CCS: 0.6t CO<sub>2</sub> (captured and stored).
- Voluntary additional information on released GHG emissions: 0.4t (process) and 0.1 t (treatment).

### Carbon Capture and Utilization

Standards for Product LCAs are currently not harmonized and do not fully address the steering effect of PCFs for important technologies with the potential to de-fossilize the chemical industry, such as carbon capture and utilization and chemical recycling. Thus, the following methodologies are a proposal by the chemical industry to steer for those technologies, but are not harmonized yet with all standards, including the GHG Protocol standard.

Captured  $\mathrm{CO}_2$  is a product of human transformation, consequently  $\mathrm{CO}_2$  is a technical flow and a chemical feedstock for  $\mathrm{CO}_2$  utilization. When  $\mathrm{CO}_2$  is captured and used, ISO- and TfS allocation hierarchy shall be used, meaning that if the multi-output situation cannot be avoided by subdivision, a system expansion or allocation following the approach described in published and accepted product category rules (PCR) or Industry Association projects, where available, for corresponding product systems, shall be applied (see chapter 5.2.9). Both approaches would not be favorable for such a new technology, as shown and explained in the following chapter. Thus, TfS decided to set up an alternative approach to be further discussed and considered.

Sources of  $\mathrm{CO}_2$  can be either direct air capture (DAC) or point sources (industrial processes like ammonia production). For both sources, the technological transition by capturing and using  $\mathrm{CO}_2$  for chemical products is driven by  $\mathrm{CO}_2$  users; for point sources, it is also driven by the  $\mathrm{CO}_2$  source. The decision-making of  $\mathrm{CO}_2$  sources and  $\mathrm{CO}_2$  users can only be steered through an assessment methodology that reflects the relationship between main product and the  $\mathrm{CO}_2$ .

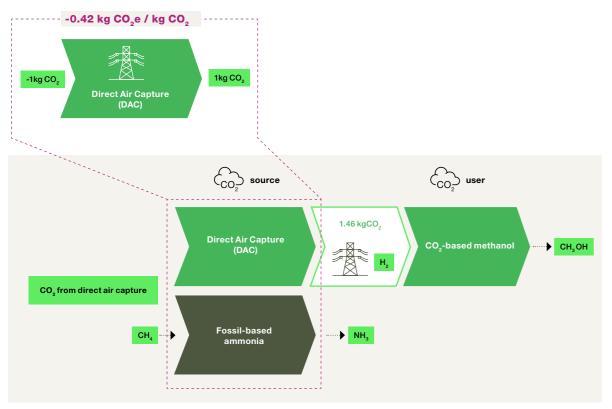
Based on exemplary data as provided in table 5.14, the impacts of different assessment methodologies were calculated for an ammonia plant (as point source) and a methanol production (user of CO<sub>2</sub>) as well as a reference CO<sub>2</sub> source from Direct Air Capture (DAC).



Figure 5.21 reflects the results of this cradle-to-gate calculation for  $\mathrm{CO}_2$ -based methanol with two different sources of  $\mathrm{CO}_2$ : Direct Air Capture (DAC) and an industrial point source, an ammonia plant. In the exemplary calculation for DAC (see Figure 5.21 and Table 5.14), 1 kg of  $\mathrm{CO}_2$  is captured via the DAC. The total PCF related to the  $\mathrm{CO}_2$  capturing and the capture process (2.52 MJ electricity per kg of  $\mathrm{CO}_2$  and 11.9 MJ heat per kg of  $\mathrm{CO}_2$ ) for DAC accounts 0.58 kg  $\mathrm{CO}_2$ e/kg  $\mathrm{CO}_2$ . Including the removal of the 1kg  $\mathrm{CO}_2$ , the total PCF for captured  $\mathrm{CO}_2$  is -0.42 kg  $\mathrm{CO}_2$ e per kg  $\mathrm{CO}_2$ .

In scenario 1) "No multi-output system", the methanol process uses 1.46 kg CO<sub>2</sub> from the DAC to produce 1kg CH<sub>3</sub>OH. As shown in Figure 5.21, accounting for all emissions from methanol and the raw material H<sub>2</sub> production, the total PCF of the CO<sub>2</sub>-based Methanol from DAC accounts 2.53 kg CO<sub>2</sub>e per kg CH<sub>3</sub>OH. Where DAC is used and thus, no capturing takes place at the ammonia plant, the production of ammonia leads to a PCF of 1.98 kg CO<sub>2</sub>e per kg NH<sub>3</sub> (Table 5.14).

Figure 5.21 DAC scenario – The CO<sub>2</sub> is captured in DAC and processed to Methanol. No CO<sub>2</sub> capturing at the ammonia plant (Table 5.15 column 1)



1kg NH<sub>3</sub> + 1 kg CH<sub>3</sub>OH = 4.51 kg CO<sub>2</sub>e

Table 5.14 PCF results using the "No-multi output" approach

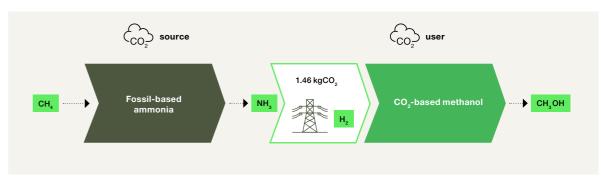
Unit: kg CO <sub>2</sub> eq	PCF ammonia as	PCF Methanol as	System expansion
	CO <sub>2</sub> source	CO <sub>2</sub> user	(1kg NH <sub>3</sub> +
	(1kg NH <sub>3</sub> )	(1kg CH <sub>3</sub> OH)	1kg CH <sub>3</sub> OH)
CO <sub>2</sub> from DAC	1.98 (no capture)	2.53	4.51

In the case where  $\mathrm{CO}_2$  is captured in a point source, the assessment approach influences the PCFs of the ammonia and the methanol. In the following example calculation, the two assessment approaches "system expansion with subsequent substitution of an ammonia plant" and "economic allocation" have been applied to show the impact of different approaches on the PCFs of the two products and thus, the steering effect.

In the scenario 2) "System expansion: avoided ammonia production w/o capture", a credit for avoided operation of an ammonia plant without capture is used to determine the PCF of captured CO<sub>2</sub> from an ammonia plant. By this approach, the 1.46 kg CO<sub>2</sub>

(needed for production of 1 kg methanol) would leave the Ammonia plant with a PCF of -0.97 kg CO $_2$  per kg CO $_2$  captured (-Avoided CO $_2$  + Emissions from capture = -1 kg CO $_2$  eq per kg + 0.03 kg CO $_2$  per kg CO $_2$ ). The PCF incentivizes the usage of CO $_2$  (PCF of CCU-Methanol accounts 1.73 kg CO $_2$ e/kg taking into consideration the negative PCF for CO $_2$ ); but does not incentivize the producer of the CO $_2$ , the fossil process of ammonia production. The ammonia production with capture would result in a PCF of 1.98 kg CO $_2$ e/kg NH $_3$  like in the first scenario of no capture.

Figure 5.22 CCU from point sources – effects of two different allocation schemes on PCF of Ammonia and PCF of Methanol (Table 5.15 column 2 and 3)



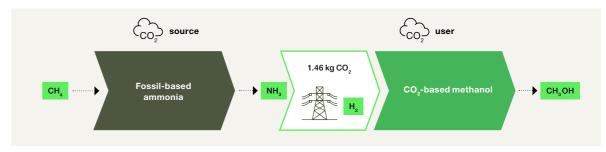
 $1 \text{kg NH}_3 + 1 \text{ kg CH}_3 \text{OH} = 3.71 \text{ kg CO}_2 \text{e}$ 

To avoid this fossil-lock in and to share incentives between both processes (CO $_2$  source and CO $_2$  user), TfS is in favor of applying system expansion with substitution of direct air capture (Table 5.15 column 4). As shown in the picture below, the ammonia producer can lower the PCF of NH $_3$  to 1.17 kg CO $_2$ e/kg NH $_3$ . The PCF of Methanol would account to 2.53 kg CO $_2$ e/kg CH $_3$ OH, which is comparable to the DAC scenario and in the average range between system

expansion with substitution of an ammonia plant (1.73 kg  $\rm CO_2e/kg~CH_3OH$ ) and economic allocation (3.25 kg  $\rm CO_2e/kg~CH_3OH$ ). The overall emission of the system is conserved 3.71 kg  $\rm CO_2e/(kg~CH_3OH+kg~NH_3)$  as for the other approaches. There is a split of incentives between the two products. The approach and rationale of system expansion with substitution of direct air capture is described in the next chapter.

Figure 5.23 CCU from point sources – proposed allocation methodology:

System expansion with avoided direct air capture (Table 5.14 column 2 and 3)



1kg NH<sub>3</sub> + 1 kg CH<sub>3</sub>OH = 3.71 kg CO<sub>2</sub>e



As  $\mathrm{CO}_2$  point sources due to their higher efficiency and high  $\mathrm{CO}_2$  concentrations are in general less emission-intensive then the energy-intensive direct air captures. The assessment approach for this specific topic should steer  $\mathrm{CO}_2$  demand towards  $\mathrm{CO}_2$  supply with minimal emissions (from point sources). Thus,  $\mathrm{CO}_2$  usage from point sources should benefit from avoiding direct air capture by applying system expansion with upstream substitution of the alternative  $\mathrm{CO}_2$  source DAC. Accordingly, the separation of  $\mathrm{CO}_2$  from  $\mathrm{CO}_2$  point sources and subsequent use in chemicals production should benefit as well because otherwise the  $\mathrm{CO}_2$  will be emitted or transferred to storage. The  $\mathrm{CO}_2$  point source (e.g. ammonia plant) and the  $\mathrm{CO}_2$  user would apply the following calculation logic:

The  $\mathrm{CO}_2$  user (here: Methanol plant) applies the PCF of  $\mathrm{CO}_2$  of using the best-in-class direct air capture process operated in the region of the point source (in this example: PCF  $\mathrm{CO}_2$  = -0.42 kg  $\mathrm{CO}_2$  e/kg  $\mathrm{CO}_2$  for DAC source). In Table 5.14 column 4, this credit is recalculated with the amount of  $\mathrm{CO}_2$  used (1.46 kg  $\mathrm{CO}_2$ ) and thus accounts -0.61 kg  $\mathrm{CO}_2$  e for the methanol production.

The  $\mathrm{CO}_2$  source (here: Ammonia plant) gets the credit of being a more efficient  $\mathrm{CO}_2$  capture than the direct air capture process. The PCF for the Ammonia would be reduced by the credit of  $\mathrm{CO}_2$  derived from the point source (compared to DAC)., i.e., here:

PCF Ammonia with capture/avoided DAC = Sum emissions Ammonia plant – (PCF avoided DAC \*  $\rm CO_2$  -Output Ammonia plant).

In the example this equals

Total per 1 kg  $NH_3$  + 1.46 kg  $CO_2$ : Emissions from Ammonia: 1.58 kg  $CO_2$  - 1.46 kg  $CO_2$  = 0.12 kg  $CO_2$ PCF Ammonia = (0.36 + 0.08 + 0.12) – (-0.61 kg  $CO_2$ e) = 1.17 kg  $CO_2$ e per kg  $NH_3$ 

Total per 1 kg CH<sub>3</sub>OH: PCF Methanol = (2.94 + 0.09 + 0.12) + (-0.61 kg CO<sub>2</sub>e) = 2.54 kg CO<sub>2</sub>e per kg CH<sub>2</sub>OH

	ummary overview of s for calculation	1) No multi- output system	2) System expansion: avoided ammonia production w/o capture	3) Allocation based on economic value	4) Point source Ammonia plant (avoided DAC)
			in	kg CO <sub>2</sub> e	
Direct Air	CO <sub>2</sub> captured (kg CO <sub>2</sub> )	1.00	-	-	-
capture	Contribution electricity direct	0.20	-	-	-
	Contribution electricity for heat	0.38	-	-	-
	Per kg CO <sub>2</sub>	-0.42 <sup>1</sup>	-	-	-
Ammonia	CO <sub>2</sub> captured (in kg CO <sub>2</sub> )	0	1.46	1.46	1.46
production	PCF of captured CO <sub>2</sub> (per kg CO <sub>2</sub> )		-0.97	0.07	-0.42
	Contribution captured CO <sub>2</sub> (per 1.46 kg CO <sub>2</sub> )	-	-	-	-0.61
	Contribution raw material production	0.36	0.36	0.29 <sup>2</sup>	0.36
	Contribution electricity consumption	0.04	0.04	0.072	0.08
	Contribution direct emissions	1.58	1.58	0.10 <sup>2</sup>	0.12
	Outputs	1 kg NH <sub>3</sub>		1kg NH <sub>3</sub> + 1.46kg CC	)2
	per outputs	1.98	1.98	0.462	1.17
CO <sub>2</sub> -based	Input CO <sub>2</sub>	1.46	1.46	1.46	1.46
Methanol production	PCF of captured CO <sub>2</sub> (kg CO <sub>2</sub> eq / (kg CO <sub>2</sub> ))	-0.42	-0.97	0.072	-0.42
	Contribution raw material production - CO <sub>2</sub>	-0.61	-1.42	0.12	-0.61
	Contribution raw material production - H <sub>2</sub>	2.94	2.94	2.94	2.94
	Contribution direct emissions	0.09	0.09	0.09	0.09
	Contribution energy consumption	0.12	0.12	0.12	0.12
	Outputs		1 kg CH <sub>3</sub> OH		
	per output	2.53	1.73	3.25	2.54
	per (1 kg NH <sub>3</sub> + 1 kg CH <sub>3</sub> OH)	4.51	3.71	3.71	3.71

The PCF for CCU processes calculated in accordance with this TfS guideline shall use system expansion with avoided DAC. For the calculation of the PCF of captured  $\rm CO_2$ , the following energy demand should use 2.5 MJ electricity per kg  $\rm CO_2$  captured additionally, 2.2 MJ of electricity are used for the provision of low temperature heat (>100°C) via a heat pump [Deutz 2021]. Note that in the examples above a PCF of electricity of 80 g  $\rm CO_2$  e per MJ of electricity were used and shall be adopted according to the specific emission factors of the electricity.

It must be noted that in LCA databases, life cycle inventories of multi output systems are often modelled differently following deviating allocation principles (e.g., physical allocation versus economic allocation) or system expansion followed by substitution. Therefore, when selecting secondary data sets, care must be taken to ensure that they comply with the allocation principles as defined in this guideline. If these are not available, they must be developed with the database providers, if possible, to achieve harmonization. Otherwise, the result of the PCF calculation also depends on whether a supplier-specific data set from a chemical company that adheres to the principles of this guideline, or a secondary data set was used. The modeling approach for systems with CCU (system extension with avoided direct air capture) suggested in the guideline is not yet reflected in the existing LCA databases with the consequence of creation of different results.

A meaningful reporting that shows the calculation approach shall be linked to the figures also for comparison of different PCF of CCU products.

### 5.2.10.5 Calculation of mass-balanced products

Mass balance is a chain of custody used in multiple industries in which it is not practical to maintain physical segregation of alternative and conventional materials during processing. Mass balance helps enable a transition to a sustainable and circular economy by enabling the efficient co-processing of alternative materials in existing large-scale assets and complex supply chains. Alternative materials can be e.g., bio-based feedstocks but also other feedstocks such as chemically recycled feedstocks, waste feedstocks or CO<sub>2</sub>-based materials. Mass balance is especially important to many companies in the chemical industry who are transitioning to use of waste plastic and bio-based materials as feedstocks to reduce the usage of virgin fossil-based materials and to help solve the global plastic waste dilemma with molecular recycling. Mass balance does require a physical link between input and outputs materials and is therefore different from more indirect chain of custody approaches such as Book and Claim.

Mass balance ensures that the quantity of output material is balanced with (does not exceed) the input of material and is appropriately adjusted for yields and conversion factors.

Co-processing of alternative and conventional raw materials results in the production of materials of mixed origin which are not distinguishable in terms of composition or technical properties. Mass balance allows alternative content to be attributed to individual outputs to create value from the use of alternative inputs. Large integrated assets cannot be transitioned immediately, and mass balance provides a critical bridge.

The mathematical approach to calculating PCF for processes in which mass balance attribution occurs is beyond the scope of this guideline because it is different for different types of chemical processes. An industry guidance, product category rule, or international standard is needed for implementation of mass balance in LCA. It is not possible to standardize an approach in this TfS guideline for this complex and emerging topic. Further standards development on chain of custody considerations in LCA is needed.

The following requirements shall apply for the usage of mass balance chain of custody in determination of PCF:

- The mass balance shall follow a transparent certification standard and the conformance to the certification shall be verified by an independent and qualified third party.
- a. The certification system shall have clear chain of custody rules, traceability requirements, defined boundaries, guidelines for marketing claims, include safeguards against double-counting, and shall identify the type of sustainable raw material throughout the supply chain.
- The LCA of manufacturing process in which the mass balance attribution occurs shall be in conformance with ISO 14044 [ISO 14044: 2006] The study shall document how the material flow and attributions were calculated.

For bio- or bio-circular attributed raw materials the biogenic uptake can be considered, but double counting shall be avoided. (e.g. biogenic uptake has to be allocated in a stoichiometric way to bio-based material and potential bio-waste streams). Therefore, high attention is necessary when allocating biogenic or bio-attributed carbon. To also reflect mass-balance products the term biogenic carbon content should be enlarged to biogenic carbon content/ attributed biogenic carbon (acc. to the mass-balance approach)

As one published example, Jeswani [Jeswani et al [2019]] described a methodology for integrating the mass balance approach into LCA for biomass applications in the chemical sector using pyrolysis followed by steam cracking. The concept conforms to the requirements ISO 14044 [ISO 14044: 2006] and may be applied to mass balance applications using bio-based feedstocks (biomass balance) or recycled feedstocks (circular mass balance). The number of sustainable feedstocks required to replace the fossil inputs are calculated through material flow analysis. The life cycle inventory of outputs with attributed sustainable content (using mass balance) are determined based on relative conversion rates of the different feedstocks and chemical values of the resulting outputs.

### 5.2.11 Data quality and share of primary data

### 5.2.11.1 Share of primary data

To create visibility on the share of primary data in PCF calculations, the primary data share (PDS) in each dataset should be determined (and shared) [Pathfinder Framework (PACT powered by WBCSD)]. This can be done by calculating the proportion (%) of the total GHG impact (CO $_2$  eq) that is derived by using primary data in the cradle-to-gate system boundary (see Formula 2). See glossary for definitions of primary and secondary data.



### Formula 3: Calculation approach of the PDS

Part of PCF based on primary data 
$$(CO_2e)$$
 = PDS<sub>PCF</sub>(%)

Ideally, the share of primary data for relevant input flows obtained from upstream suppliers (tier n-1) are available. If so, the PDS of the PCF should be calculated using a PCF attributed average approach of the material and energy inputs. If not all members of the supply chain are encouraged to participate in this effort as the share of primary data can only be accurately determined if the respective information for most inputs is provided by the respective suppliers.

To do so, the individual PDS received from supplier (PDS external components) as well from other components (PDS other components), e.g., energy inputs or direct emissions from production, should be multiplied with their respective relative contribution (in %) to the PCF. All weighted PDS (weighted PDS components) should then be added up to obtain an overarching PDS (PDS output). To help increase transparency on primary data use, information on PDS should be shared downstream (tier n+1) together with the PCF. The inclusion of an explanation for the share of primary data is thus encouraged, with the objective of helping businesses support each other in increasing the amount of primary data flowing through the system and ensuring more accurate PCFs if the quality of the data is in addition very good (Figure 5.24). The general approach is shown in Figure 5.25.

Figure 5.24 Calculation of Primary data shares of two components

### **PDS** calculation

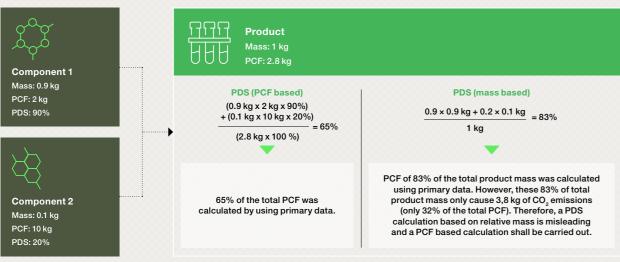
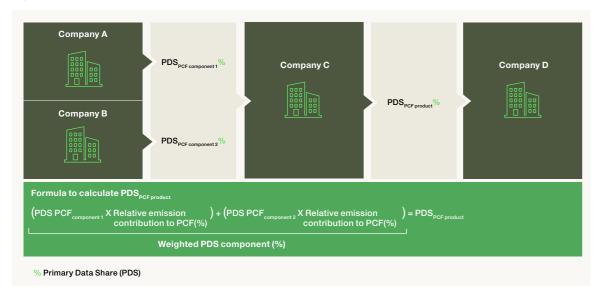


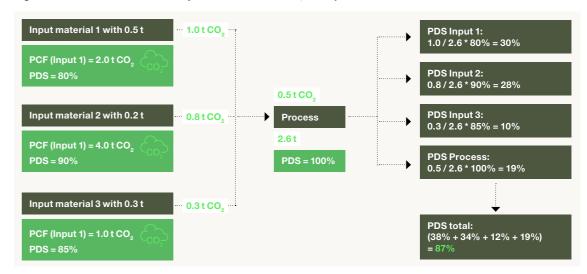
Figure 5.25 Calculation of Primary data share for a PCF [Pathfinder Framework (PACT powered by WBCSD)]



A detailed example is shown in Figure 5.26 with application of detailed steps in the generation of a PDS for a PCF of a product. A primary data share shall only be attributed, if both the activity data (e.g. amount in kWh) and the emission

factor information is derived from primary sources. If one of these two information is derived from secondary data, the whole PDS for this unit process is rated as secondary data.

Figure 5.26 Calculation of Primary data share for a PCF, example



In general, a primary dataset can only be rated as primary, if the activity data (data source) and the emission factor (EF source) is a primary information. If only one of the two factors of the PDS calculation is secondary, the

whole rating for this contribution is rated as secondary and will be implemented in the PDS calculation accordingly. An example is given in Table 5.15.

Table 5.16 PDS calculation example for primary and secondary data sources

Material	Data input (kWh)	Data source	EF (kg CO <sub>2</sub> e)	EF source	kg CO₂e	% PCF	Total PDS
A	10,435	Primary	0.19	Primary	1,983	42%	42%
В	10,000	Secondary	0.18	Secondary	1,800	38%	0%
С	5,000	Primary	0.18	Secondary	900	19%	0%
					4,683		42%

### 5.2.11.2 Data quality rating (DQR)

During the data collection process, companies shall assess the data quality of activity data, emission factors, and/or direct emissions data by using the data quality indicators.

If higher-quality data exists in-house than available in secondary databases (for example, in-house emission factors for fuel) and is used for calculations, the adequacy of such in-house data shall be reviewed and reported in a DQR following the criteria outlined in this chapter. Data sourced from verified emission factor databases (See chapter 5.2.6) shall be reported in a DQR as well, addressing its representativeness, relevance, and correct application to the product in question as well. The calculation and reporting of a DQR becomes mandatory only from 2025 on to give companies enough time to prepare for this. Until then it is recommended to do it on a voluntary basis.

The standard defines five data quality indicators to use in assessing data quality. They are shown below and summarized in Table 5.16.

 Technological representativeness: the degree to which the data reflect the actual technology(ies) used in the process.

- Geographical representativeness: the degree to which the data reflects actual geographic location of the processes within the inventory boundary (e.g., country or site).
- Temporal representativeness: the degree to which the data reflect the actual time (e.g., year) or age of the process.
- Completeness: the degree to which the data are statistically representative of the process sites.
- Reliability: the degree to which the sources, data collection methods, and verification procedures used to obtain the data are dependable.

Assessing data quality during data collection allows companies to make data quality improvements more efficiently than when data quality is assessed after the collection is complete.

The Pathfinder Framework requires only those inputs representing more than 5% of the total PCF to undergo the DQR assessment which reduces the workload for the generation of DQR factors. TfS recommends this approach as well (Table 5.17).

47

46 South a detail of the contract with and the contract with a contract with a



Table 5.17 Data quality indicators of GHG Protocol

Indicator	Description	Relation to data quality
Technological representativeness	The degree to which the data reflects the actual technology(ies) used.	Companies should select data that are technologically specific.
Temporal representativeness	The degree to which the data reflects the actual time (e.g., year) or age of the activity.	Companies should select data that are temporally specific.
Geographical representativeness	The degree to which the data reflects the actual geographic location of the activity (e.g., country or site).	Companies should select data that are geographically specific.
Completeness	The degree to which the data are statistically representative of the relevant activity.  Completeness includes the percentage of locations for which data is available and used out of the total number that relate to a specific activity. Complete- ness also addresses seasonal and other normal fluctuations in data.	Companies should select data that are complete.
Reliability	The degree to which the sources, data collection methods and verification procedures <sup>1,2</sup> used to obtain the data are dependable.	Companies should select data that are reliable.

Table 5.18 Data quality assessment used in TfS and [Pathfinder Framework (PACT powered by WBCSD)]

DQI	1 - Good	2 - Fair	3 - Poor
Technology	Same technology	Similar technology (based on secondary data)	Different or unknown technology
Time	Data from reporting year	Data less than 5 years old	Data more than 5 years
Geography	Same country or country subdivision	Same region or subregion	Global or unknown
Completeness	All relevant sites for specified period	<50% of sites for specified period or >50% of sites for shorter period	Less than 50% of sites for shorter time period or unknown
Reliability	Measured activity data	Activity data partly based on assumptions	Non-qualified estimate

The quality assessment of data based on the Table 5.14 can be used to derive more quantitative information in form of a Data Quality Rating (DQR) to give users of the data a better impression of the overall quality of data and the resulting PCF.

The data quality of each PCF shall be calculated and reported. The DQR calculation shall be based on five data quality criteria (each criterion considered of equal importance) where TeR is the Technological-Representativeness, TiR is the Time/Temporal

Representativeness, GeR is the Geographical-Representativeness, C is completeness and R is reliability.

The quality levels are expressed in four categories from, 1 Good, 2 Fair, 3 Poor. The representativeness (technological, geographical, and time-related) characterizes the degree to which the processes and products selected depict the system analyzed, while the completeness and reliability addresses the quality of the generated PCF result.

The contributions of the input materials to the PCF of the process are linked with the DQR of the input material. As lower the DQR score and as higher the share of the total PCF for an input material is, as more positive is the impact of an input material to the overall DQR process score.

For example:	Product 1	Product 2
Technological	2	3
representativeness (TeR):		
Temporal	1	3
representativeness (TiR):		
Geographical	2	2
representativeness (GeR):		
Completeness (C):	3	3
Reliability (R):	2	3
Total	10	14
DQR total (Total / 5)	2	3

In Formula 4, the aggregation of all single results of all input materials from upstream is shown. The second line shows the general calculation of a single DQR of a process, based on the five criteria described above. In line 3 it is shown, how to add process related DQR and the upstream DQR according to Figure 5.27.

### Formula 4: General calculation of data quality ratings

DQR<sub>unstream</sub> = (DQR Inputmaterial 1\* PCF<sub>total</sub> share 1+ DQR Inputmaterial 2\* PCF<sub>total</sub> share 2 + DQR Inputmaterial 3\* PCF<sub>total</sub> share 3+ DQR Inputmaterial n\* PCF<sub>total</sub> share n)

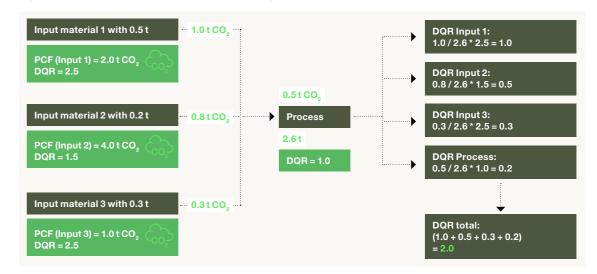
$$DQR_{process} = (TeR + TiR + GeR + C + R) / 5$$

$$DQRtotal = (DQR_{upstream} + DQR_{process})$$

The DQR<sub>process</sub> shall be calculated for the output of e.g. 1kg or 1t output as defined in the declared unit.

For an example see Figure 5.27. The Total DQR for this process is 2.0 and shall be reported to the recipient of the PCF data as well after 2025. The DQR can be used as an input for complete LCA which enables the final calculation of a complete DQR. The DQR supports the interpretation of PCF data and supports the identification of improvement potentials of the quality of the PCF data. Note that the DQR of the process is not always equal to 1 depending on which data are available. A process where primary data can be generated, can always have a low Technological or Temporal representativeness ending in lower scores than 1. Company owned processes must be assessed in the same manner as upstream processes.

Figure 5.27 DQR example for a process including upstream DQR (Source TfS)



### Improving data quality

Collecting data and assessing its quality is an iterative process for improving the overall data quality of the product inventory. If data sources are identified as low quality using the data quality indicators, companies should re-collect data [GHG Protocol Product Standard].

The following steps are useful when improving data quality:

1. Identify sources of low-quality data in the product inventory using the data quality assessment results. Sources with low quality data that have been identified as significant to the PCF results should be given priority.

- 2. Collect new data for the low-quality data sources as resources allow.
- 3. Evaluate the new data. If it is of higher quality than the original data, use in its place. If the data are not of higher quality, either use the existing data or collect new data.
- 4. Repeat as necessary and as resources allow. If companies change data sources in subsequent inventories, they should evaluate whether this change creates the need to update the base inventory.

<sup>(1)</sup> Adapted from B.P. Weidema, and M.S. Wesnaes, "Data quality management for life cycle inventories - an example of using data quality indicators," Journal of Cleaner Production. 4 no. 3-4 (1996): 167-174.

(2) Verified data: Verification may take place in several ways, e.g., by on-site checking, by recalculation, through mass balance, or by crosschecks with other sources.



### 5.3 Verification and Reporting

Table 5.19 Reporting examples in different approaches of companies

Example case	Applicable PCF calculation rule	Voluntary additional information for emission offset
The company purchases emission credits from a project investing in reforestation to offset 50% of the calculated PCF	The PCF remains the same as calculated.	The emissions offset of 50% may be provided separately from inventory results
The company purchases emission credits from a carbon capture and storage facility to offset 30% of the calculated PCF	The PCF remains the same as calculated.	The emissions offset of 30% may be provided separately from the inventory results
The company purchases renewable electricity certificates to offset 100% of the electricity consumption of a particular site, and as a consequence, reduces to zero the electricity-related emissions of the PCF	The PCF is reduced according to the reduction potential of electricity use Offsets are not taken into account as credits	The emission offset may be provided separately from inventory results
The company generates direct CO <sub>2</sub> within a reaction, which is captured and sold as a by-product. (see Chapter 5.2.10.4)	The impact of the process capturing atmospheric CO <sub>2</sub> and sold as a by-product shall be added to the inventory results of the PCF according to the amount of CO <sub>2</sub> captured, and may be subtracted from the inventory results of the process	As an alternative to subtracting the CO <sub>2</sub> emissions captured and sold from the inventory results, the emissions captured may also be provided separately

### 5.3.1 Verification of PCF calculations / Quality Assurance

A **verification** is defined as the confirmation, through the provision of objective evidence, that specified requirements have been fulfilled [ISO 9000: 2005]. To achieve the verification, the calculations shall be cross-checked against the requirements of this guideline and the results shall be reported.

Verification of PCF data prior to sharing is strongly encouraged to ensure high quality and trustworthy data [Pathfinder Framework (PACT powered by WBCSD)]. A significant update has been made to the verification and assurance chapter within the Pathfinder Framework.

No verification is not allowed under TfS. Possible types of verification can be an internal LCA expert, a third party verification - product review or a independent party verification - systematic approach review. The type of verification must be reported together with the PCF (see 5.3.2).

If internal verification is done, the claim of the PCF figure shall always include that the PCF was calculated in alignment with the TfS Guideline, while a third-party verification allows a stronger claim (e.g. verified against TfS)

Any type of verification should include a 4-eye-principal check by an internal LCA expert or external auditor regarding the following aspects:

- The goal and scope and its related aspects (see 5.1).
- The calculation rules (see 5.2).
- The system boundaries (see 5.1.2 and cut-off criteria (see 5.2.3).
- The data quality (see 5.2.5).

A third-party verification can be helpful, to fulfill these requirements. This can either be done on a product level or in a systematic approach verification where a company methodology to calculate consistent PCF is verified.

The **quality assurance** is defined as part of quality management focused on providing confidence that quality requirements will be fulfilled. In this sense, a quality assurance shall address, if the PCF results and the approach to achieve them fulfills requirements of high quality beyond data quality (adopted from [ISO 9000: 2005]).

The following short checklist can help the LCA practitioner to validate the PCF. Besides the LCA expert, people that can support the validation, include technology experts, controllers, plant managers and site managers:

- Check the overall mass balance (includes raw material inputs, product outputs, wastes as well as emissions into air and water).
- · Completeness of life cycle stages.
- Check the elementary balance by doing a stochiometric calculation.
- Check if direct emissions are realistic, e.g., by carbon balance.
- Check if the carbon balance is closed, all inputs are considered and balanced with outputs to products, emissions (air, water, soil), wastes. Check if process related direct emissions are plausible (carbon, nitrogen process input output balances).
- Check data aggregation, data polishing and underlying modeling to calculate product inventory of your own data sets.
- Check if correct calculation formulas where applied.
- Check utility consumption (plausible?).
- Check allocation factors (in line with chapter 5.2.9): the sum of the allocated inputs and outputs of a unit process equals the inputs and outputs of the unit process before allocation and allocation factors over all co-products of one multi-output process sum up to 1.
- CO<sub>2</sub>e benchmark against own calculations, same product from other sites/plants companies, existing LCA data, LCIs from other third-party databases.
- Check if biogenic emissions and uptakes are correctly considered and reported (5.2.10.1).
- Check the appropriateness of the secondary datasets selected for Scope 3 upstream data:
- Check if technology and geography represented in the LCI is the appropriate.
- Check if the application of proxies is appropriate
- If supplier data is available replace dataset.
- Check if a data quality score was generated and if it is meaningful.
- Check why there are significant deviations to LCA benchmark data.
- Sensitivity analysis and quality checks of results: Sensitivity analyses with different modeling choices (e.g., another dataset for a raw material, another allocation method for the foreground product system) should be performed to test the robustness of the result.

Any additional information available such as a PCF report or a critical review statement can be added or attached to complement and provide more details to the information [BASF SE (2021)].

Results reported in the PCF study report may be used in footprint communications [ISO 14026: 2017].

### 5.3.2 Information to be reported with PCF

This section specifies the information requirements to be provided by suppliers alongside PCF values. Additional information besides the PCF value is needed to support the interpretation and validation of PCF data, as well as to provide necessary information for quantification of customer PCFs further down the value chain. The PCF covers one environmental impact. In this context it should be mentioned, that no overall statements on the environmental performance of the product can be given. Comparisons of PCF are only possible under certain criteria if all relevant information is reported.

The fields marked as "mandatory" in the table 5.19 ("yes") shall be provided by suppliers when disclosing PCF values. Some fields will become mandatory from 2025 onwards to provide a transition period for adaptation. TfS still highly recommends reporting as much data as possible. Additional details, currently not mandatory, may also be provided if available to provide further support. ISO 14067 [ISO 14067: 2018] describes requirements for reporting which are reflected in the attributes list. To be fully compliant for a PCF study, all reporting requirements shall be addressed. In the B2B exchange, if no further information is requested, the following GHG values shall be the basis for a PCF study report:

- Declared unit.
- Total GHG emissions and removals. Optional their link to the main life cycle stages in which they occur, including the absolute and the relative contribution of each life cycle stage.
- Net fossil GHG emissions and removals.
- Biogenic GHG emissions and removals.
- GHG emissions and removals occurring because of land use and direct land use change.
- Biogenic carbon content of products.
- Functional or declared unit and reference flow.
- The selected cut-off criteria and cut-offs.
- The selected allocation procedures.
- Description of data and data quality.
- Treatment- and use of electricity.
- Description of the stages of the life cycle.
- Time period for which the PCF is representative.
- A graphical presentation of results of the PCF.
- Where an aviation multiplier is used, the effect of this multiplier shall not be included in the PCF and shall be reported separately together with the source.



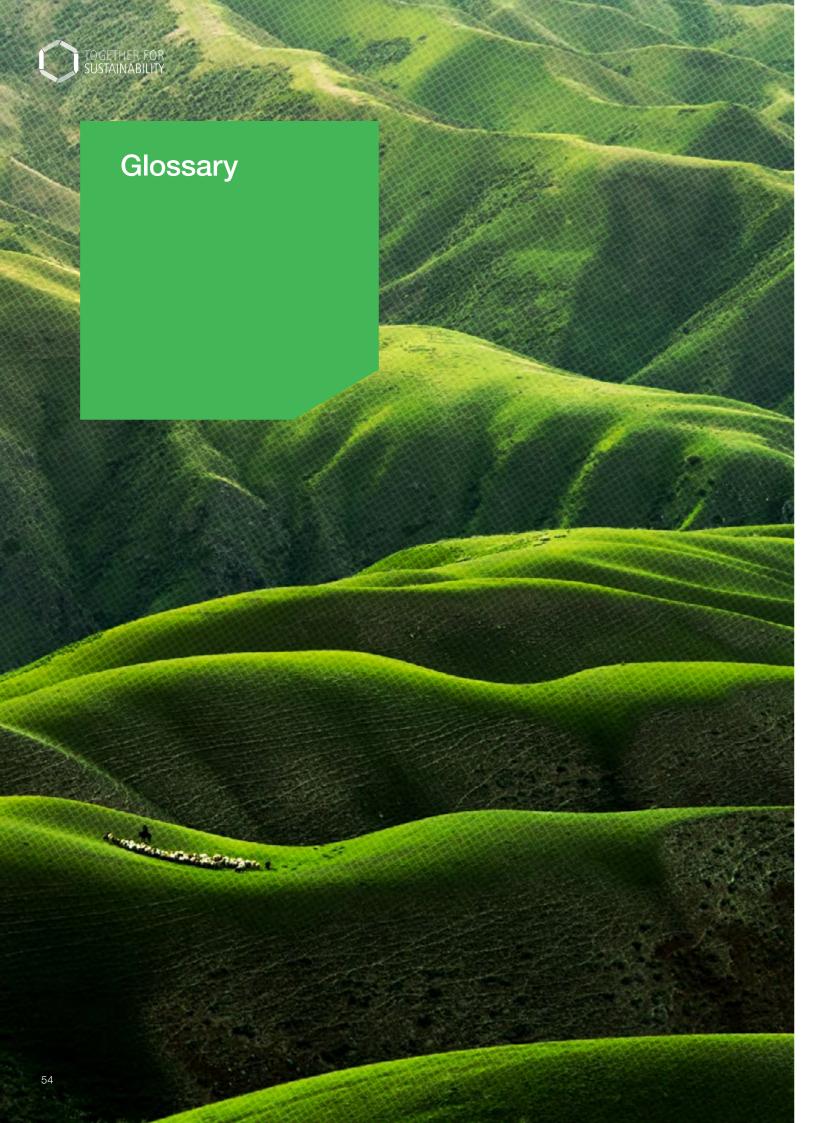
**Table 5.20** 

14016 3.20				
Category	Attribute	Further explanation (Semantics and ILCD)	Example	Mandatory
Company and product	Company name	(Legal) Name of Data owner	My corp	yes
	Company ID	Abstract ID	E.g. # 311	yes
	Product trade name	Product name	Green Ethanol	yes
	CAS	CAS Number	58-08-2	yes, if available
	Declared unit: kg	Unit of analysis of the product (always kg)	1 kg	yes
	Product description including reference to the solution for which PCF is reflected	(Technical) description of product or waste	Ethanol, 95% solution	yes
PCF	Unique ID	ID of exchanged PCF, e.g. UUID	58-08-2-0017	no
	PCF (excl. Biogenic emissions and removals)	Cradle-to-gate PCF in kg CO <sub>2</sub> e/kg product Sum of separate emission values 1 + 2 + 3	2.6 kg CO <sub>2</sub> e/kg Ethanol	yes
	PCF (incl. biogenic emissions and removals)	Cradle-to-gate PCF in kg CO <sub>2</sub> e/kg product Sum of separate emission values 1+2+3+4	0.7 kgCO <sub>2</sub> e/kg Ethanol	yes, from 2025 on
	Separated into emission values:  1. Fossil CO <sub>2</sub> e-emissions (net result of fossil emissions and removals)  2. Biogenic CO <sub>2</sub> e-emissions (only other GHG emissions than CO <sub>2</sub> – excludes biogenic CO <sub>2</sub> )  3. Land use and direct land use change CO <sub>2</sub> e-emissions  4. Biogenic removals (biogenic CO <sub>2</sub> contained in the product)	In kg CO <sub>2</sub> e/kg product	1. Fossil CO <sub>2</sub> e: 2.0 kg CO <sub>2</sub> e/kg Ethanol 2. Biogenic CO <sub>2</sub> e*: 0.4 kg CO <sub>2</sub> e/kg Ethanol 3. Land use /LUC CO <sub>2</sub> e: 0.2 4. Biogenic removal: -1.9 kg CO <sub>2</sub> e/kg Ethanol	Separated emission values: yes, from 2025 (at least 1, 2-4 only if applicable) Please keep in mind that reporting is mandatory if a compliance with ISO 14067, PEF or the Pathfinder is anticipated
		Kg C/kg product	0.521 kg/kg Ethanol	yes
	Reference period (year or start/end date if > one yr) and version (if revised within reference period)	Year/period of PCF calculation	2021, v 2.0 Or 01/01/2020 – 31/12/2021	yes
	Geography (as specific as possible)	Location of production / product	Global, Europe, Germany, or Ludwigshafen, 67063, Germany	yes

 $<sup>^{\</sup>star}$  If the share of biogenic CO $_2$  emissions is not known and cannot be determined, the calculated CO $_2$  emissions shall be considered as fossil CO $_2$  emission. In this case the CO $_2$  removal shall only be calculated based on the carbon content in the product.

Category	Attribute	Further explanation (Semantics and ILCD)	Example	Mandatory
	Technological reference *	Technological description	Electrolysis	yes, from 2025 on
	Data quality rating (DQR)	DQR in score from 1 to 3	DQR 1.5	yes, from 2025 on
	Primary data share (PDS)	PDS in %	PDS 95%	yes, from 2025 on Please keep in mind that reporting is mandatory if compliance with Pathfinder is anticipated
	Source of secondary data and version	Refers only to source of secondary data at reporting company	ILCD, Carbon Minds, ecoinvent 3.8, open sources	yes
	Allocation method used*	Type of allocation	Mass allocation	yes
	Verification approach (None, Internal LCA Expert, Third Party Verification - Product Review, Third Party Verification - Systematic Approach Review)		Verification by internal LCA expert	yes
Boundary & standards	PCF calculation Standards or guidelines used (or product or sector specific rules if used)	Standard used for calculating the PCF	PCR, TfS Guideline 2022, ISO 14067: 2018	yes
Additional information – biobased materials	Biogenic carbon content (physical or BMB)	Kg Bio-C/kg product	0.52 kg biogenic C/kg Ethanol	yes, from 2025 on
Additional information – waste incineration	Allocation approach used for waste incineration with energy recovery	Cut-off, reverse cut-off or system expansion		yes
Additional information – chemical	Recycled carbon content (physical or BMB)	Kg recycled-C/kg product	0.5 kg recycled C /kg Ethanol	no
recycled material	Allocation method used for recycled carbon content		Upstream System expansion or cut-off	no
	Type of recycled content		Post-industrial, post-consumer	no
Additional information	CCU-based carbon content	Kg CCU-C/kg product	0.5 kg recycled C/kg Ethanol	no
- Captured and used CO <sub>2</sub> material	Allocation method used for CCU		System expansion and substitution	no
2	CO <sub>2</sub> -origin	Source from where CO <sub>2</sub> is captured	DAC or Point source ammonia plant	no
Additional information – general	Further information on modelling	Assumptions and limitations	Cut-off set on 6%	no

 $<sup>^{\</sup>star}\,\text{Of}$  the foreground system, i.e. the last process in the value chain that is calculated.



Abbreviation	Term	Definition
	Activity data	"Activity data are quantified measures of a level of activity that result in GHG emissions or removals". Activity data can be measured, modeled, or calculated.
		There are two categories of activity data: process activity data and financial activity data.
		Process activity data are physical measures of a process that results in GHG emissions or removals. These data capture the physical inputs, outputs, and other metrics of the product's life cycle (e.g. energy, mass, volume etc). Financial activity data are monetary measures of a process that results in GHG emissions.  [GHG Protocol Product Standard]
	Allocation	Partitioning the input or output flows of a process or a product system between the product system under study and one or more other product systems. [ISO 2006].
	Background data	See also secondary data. Data that concern processes outside the operational control of the company.
	Bill of materials (BOM)	A structured list of all the components, and their quantities that make up an assembly or product.
	Biogenic carbon content	Fraction of carbon derived from biomass in a product.
	Biogenic emissions	CO <sub>2</sub> emissions from the combustion or biodegradation of biomass.
	Biogenic removals	The sequestration or absorption of GHG emissions from the atmosphere, which most typically occurs when $\mathrm{CO}_2$ is absorbed by biogenic materials during photosynthesis.
	Biomass	Material of biological origin excluding material embedded in geological formations and/or fossilized.
CO <sub>2</sub> e	Carbon Dioxide Equivalent	Carbon dioxide equivalent, or $\mathrm{CO_2e}$ is a metric measure representing all greenhouse gases by converting them to the equivalent amount of $\mathrm{CO_2}$ .
	Declared unit	Intermediate or final products, that is, products which will still be processed further to create a final product, can, however, have several functions based on their eventual end use. In this case (and where an LCA does not cover the full life cycle), the term declared unit – typically referring to the physical quantity of a product, for example "1 liter of liquid laundry detergent with 30 percent water content" – shall be used instead.
	Functional unit	A functional unit describes the function of a product in question. For example, for a laundry detergent, the functional unit could be defined as "washing 4.5 kg of dry fabric with the recommended dosage with medium-hard water". Understanding the functional unit is essential for comparability between products with the same function, as it provides the reference to which the input (materials and energy) and output (such as products, byproducts, waste) are quantified.
	Cradle-to-gate	An assessment that includes part of the product's life cycle, including material acquisition through the production of the studied product and excluding the use or end-of-life stages. (WRI and WBCSD 2010).
	Cradle-to-grave	A cradle to grave assessment considers impacts at each stage of a product's life cycle, from the time natural resources are extracted from the ground and processed through each subsequent stage of manufacturing, transportation, product use, recycling, and ultimately, disposal. (Athena Institute & National Renewable Energy Laboratory draft 2010).



Term	Definition
Greenhouse Gases	Greenhouse gases constitute a group of gases contributing to global warming and climate change. The Kyoto Protocol, an environmental agreement adopted by many of the parties to the United Nations Framework Convention on Climate Change (UNFCCC) in 1997 to curb global warming, nowadays covers seven greenhouse gases:  The non-fluorinated gases:  Carbon dioxide (CO <sub>2</sub> ).  Methane (CH <sub>4</sub> ).  Nitrous oxide (N <sub>2</sub> O).  The fluorinated gases:  Hydrofluorocarbons (HFCs).  Perfluorocarbons (PFCs).  Sulphur hexafluoride (SF <sub>6</sub> ).  Nitrogen trifluoride (NF <sub>3</sub> ).  Converting them to carbon dioxide (or CO <sub>2</sub> ) equivalents makes it
	possible to compare them and to determine their individual and total contributions to global warming.
Conformity assessment	Demonstration that specified requirements relating to a product, process, system, person or organization are fulfilled  Note 1 to entry: Adapted from ISO/IEC 17000: 2004, definition 2.1.  ISO/TS 14441:2013(en), 3.13
Consumption mix	This approach focuses on the domestic production and the imports taking place. These mixes can be dynamic for certain commodities (e.g., electricity) in the specific country/region.
International Organization for Standardization	
ISO standard on Greenhouse gases — Carbon footprint of products — Requirements and guidelines for quantification	ISO 14067: 2018 specifies principles, requirements and guidelines for the quantification and reporting of the carbon footprint of a product (CFP), in a manner consistent with International Standards on life cycle assessment (LCA) [ISO 14040 [ISO 14040: 2006] and ISO 14044].
Life Cycle Assessment	The compilation and evaluation of the inputs, outputs, and the potential environmental impacts of a product system throughout its life cycle [ISO 1440: 2006].
Life Cycle Inventory	The phase of life cycle assessment involving the compilation and quantification of inputs and outputs for a product throughout its life cycle [ISO 14040:2006].
Life Cycle Impact Assessment	The phase of life cycle assessment aimed at understanding and evaluating the magnitude and significance of the potential environmental impacts for a product system throughout the life cycle of the product [ISO 14040:2006].
Primary data	Sometimes also called activity data. Data that concern processes inside the operational control of the company or data from specific processes in the product life cycle.
	A partial PCF is considered primary data if the measure of the activity data and the measure of the emission factor are based on data where the data generators have a direct access to via direct measurements or assessments where they have a direct control.
	"Data pertaining to a specific product or activity within a company's value chain. Such data may take the form of activity data, emissions or emission factors. Primary data is site-specific, company-specific (if there are multiple sites for the same product) or supply chain-specific. Primary data may be obtained through meter readings, purchase records, utility bills, engineering models, direct monitoring, material or product balances, stoichiometry or other methods for obtaining data from specific processes in the value chain of the company."  Path 2021:41]
	Conformity assessment  Consumption mix  International Organization for Standardization  ISO standard on Greenhouse gases — Carbon footprint of products — Requirements and guidelines for quantification  Life Cycle Assessment  Life Cycle Inventory  Life Cycle Impact Assessment

Abbreviation	Term	Definition
PCF	Product Carbon Footprint	The Product Carbon Footprint is the most established method for determining the climate impact of a product, considering the total greenhouse gas (GHG) emissions caused to produce a product, expressed as carbon dioxide equivalent. The PCF can be assessed from cradle-to-gate (partial PCF) or from cradle-to-grave (total PCF).
PCR	Product Category Rules	Set of specific rules, requirements, and guidelines for developing Type III environmental declarations for one or more product categories. [ISO 14025:2006].
	Production mix	This approach focuses on the domestic production routes and technologies applied in the specific country/region and individually scaled according to the actual production volume of the respective production route. This mix is generally less dynamic.
	Removal	The sequestration or absorption of GHG emissions from the atmosphere, which most typically occurs when $\mathrm{CO}_2$ is absorbed by biogenic materials during photosynthesis.
GHG Protocol	Scope 1 Emissions	Scope 1 emissions include GHG that arise from the combustion of fuels owned or controlled by the reporting organization.
GHG Protocol	Scope 2 Emissions	Scope 2 emissions include GHG emissions that result from the consumption of purchased or acquired energy such as electricity, heating, cooling, and steam.
GHG Protocol	Scope 3 Emissions	Scope 3 emissions include the remainder of indirect GHG emissions which cannot be categorized as energy-related emissions in Scope 2.
	Secondary data	See also background data. Data that concern processes outside the operational control of the company or process data that are not from specific processes in the product life cycle.
		"Data that is not from specific activities within a company's value chain but from databases, based on averages, scientific reports or other sources."  Path 2021:41]
	System expansion	Expanding the product system to include the additional functions related to the co-products. System expansion is a method used to avoid co-product allocation.
	Utilities	The term "utilities" includes here: Electricity, process steam, excess steam, cooling water, demineralized water, process water, compressed air and nitrogen.
	Validation	The process of evaluating a system or component to ensure compliance with the functional, performance and interface requirements.
		ISO/IEC 14776: 2010
	Verification	Confirmation, through the provision of objective evidence, that specified requirements have been fulfilled.
		[SOURCE: ISO 9000: 2005] ISO 14025:2006
	Unit process	Smallest element considered in the life cycle inventory analysis (3.1.4.4) for which input and output data are quantified. [ISO 14040:2006], 3.34]
	Waste	Substances or objects which the holder intends or is required to dispose of.  NOTE This definition is taken from the Basel Convention on the Control of Transboundary Movements of Hazardous Wastes and Their Disposal (22 March 1989), but is not confined in this International Standard to hazardous waste.  [ISO 14040:2006], 3.35]



### References

AIB - European Residual Mix, (2022), European Residual Mixes- Results of the calculation of Residual Mixes for the calendar year 2021, https://www.aib-net.org/facts/european-residual-mix; (Accessed 18 August, 2022)

BASF SE, (2021), Guideline for Product Carbon Footprint Calculations of companies, BASF SE

Contribution of Products to Avoided GHG Emissions, (2015), Guidlines for Assessing the Contribution of Products to avoided Greenhouse Gas Products, The Institute of Life Cycle Assessment, Japan

Deutz, S., Bardow, A.; (2021), Life-cycle assessment of an industrial direct air capture process based on temperature–vacuum swing adsorption. Nat Energy 6, 203–213 (2021). https://doi.org/10.1038/s41560-020-00771-9

EC- Renewable Energy Directive , (2009), Directives on the promotion of the use of energy from renewable sources and amending and subsequently repealing Directives 2001/77/EC and 2003/30/EC.

EcoTransilT- Emission Calculator for GHG Emissions, (2020), EcoTransIT World Emissions Calculator. https://www.ecotransit.org/en/emissioncalculator/

EN 15804+Amd 2:2019, (2019), Sustainability of Construction Works- Environmental Product Declarations-Core rules for the Product Category of Construction Products

EPA- Energy Recovery from MSW, (2022), Energy Recovery from the Combustion of Municipal Solid Waste (MSW) - https://www.epa.gov/smm/energy-recovery-combustion-municipal-solid-waste-msw; (Accessed 18 August 2022).

ERASM, (2014), ERASM SLE (2014). Surfactant Life Cycle and Ecofootprinting Project; updating the life cycle inventories for commercial surfactant production. Final Report for ERASM (www.erasm.org), 186 p.

Estimating and Reporting the Comparative Emissions Impacts of Products, (2019), Estimating and Reporting the Comparative Emissions Impacts of Products. Working Paper. Washington, DC: World Resources Institute. Available online at http://www.wri.org/publication/comparative -emissions; (Accessed 18 August 2022)

EU PEF, (2021), Final Product Environmental Footprint Category Rules and Organisation Environmental Footprint Sector Rules; https://ec.europa.eu/environment/eussd/ smgp/PEFCR\_OEFSR\_en.htm; (Accessed 18 August 2022)

EU Waste Framework Directive, (2008), European Waste Framework Directive 2008/98/EC; Directive on waste and repealing certain Directives

BASF (2020), ChemCyclingTM: Environmental Evaluation by Life Cycle Assessment (LCA); https://www.basf.com/global/de/who-we-are/sustainability/we-drive-sustainable-solutions/circular-economy/mass-balance-approach/chemcycling.html, (Accessed 7 September, 2022)

Eurochlor, (2022), The Chlorine Alkali Process, Final Report

GHG Protocol Corporate Standard, (2004), The Greenhouse Gas Protocol- A Corporate Accounting and Reporting Standard

GHG Protocol Corporate Standard, (2011), Corporate Value Chain (Scope 3) Accounting and Reporting Standard

GHG Protocol Product Standard, (2011), The Greenhouse Gas Protocol - Product Life Cycle Accounting and Reporting Standard

GHG Protocol Scope 2 Guidance, (2015), GHG Protocol Scope 2 Guidance- An Amendment to the GHG Protocol Corporate Standard

GHG Protocol Technical Scope 3 Guidance, (2013), Technical Guidance for Calculating Scope 3 Emissions

GLEC Framework, (2019), Global Logistics and Emission Council Framework - Logistics Emissions Accounting and Reporting

Global Compact Scope 3.1, (2017), Practical Guidelines for Data Collection and Calculation of Greenhous Gas Emissions from Purchased Goods and Services: Global Compact Network Germany

Green-e 2021- Residual Mix Emission Rate, (2021), Residual Mix Emission Rate (2019 Data) 2021 https://www.green-e. org/2021-residual-mix; (Accessed 18 August 2022)

ICCA - Avoided Emission Challenge, (2013), Addressing the Avoided Emission Challenge, Guidelines from the chemical industry for accounting for and reporting greenhouse gas (GHG) emissions avoided along the value chain based on comparative studies

ICCA - Avoided Emission Challenge, (2017), Avoided GHG Emissions- The Essential Role of Chemicals. Accounting for and Reporting Greenhouse Gas (GHG) Emissions Avoided along the Value Chain based on Comparative Studies. Version 2 October 2013 (updated December 2017)

IPCC - The Physical Science, (2013), Climate Change 2021-The Physical Science Basis. https://www.ipcc.ch/report/ ar5/wg1/; (Accessed 18 August 2022)

IPCC - The Physical Science, (2021), Climate Change 2021-The Physical Science Basis. https://www.ipcc.ch/report/ sixth-assessment-report-working-group-i/; (Accessed 18 August 2022)

IPCC - The Supplementary Material, (2021), The Earth's Energy Budget, Climate Feedback, Climate Sensitivity-Supplementary Materials. https://www.ipcc.ch/report/ar6/wg1/downloads/report/IPCC\_AR6\_WGI\_Chapter\_07\_Supplementary\_Material.pdf; (Accessed 18 August 2022)

IPCC- GHG Inventories Guidlines, (2006), IPCC Guidlines for National Greenhouse Gas Inventories.



ISO 14025:2006, (2006), Environmental labels and declarations — Type III environmental declarations — Principles and procedures

ISO 14026:2017, (2017), Environmental labels and declarations — Principles, requirements and guidelines for communication of footprint information

ISO 14044:2006+Amd 2: 2020, (2020), Environment Management-Lifecyle Assessment- Principles and Framework

ISO 14040:2006+Amd 1: 2021, (2020), Environment Management-Lifecyle Assessment- Principles and Framework

ISO 14064 -1:2019, (2019), Treibhausgase - Teil 1: Spezifikation mit Anleitung zur quantitativen Bestimmung und Berichterstattung von Treibhausgasemissionen und Entzug von Treibhausgasen auf Organisationsebene

ISO 14064 -2:2019, (2019), Part 2: Specification with guidance for the quantitative determination and reporting of greenhouse gas emissions and removals at the organization

ISO 14064 -3:2019, (2019), Part 3: Specification with guidance for the quantitative determination and reporting of greenhouse gas emissions and removals at the organization level

ISO 14067:2018, (2018), Greenhouse Gases- Carbon Footprint for products- Requirements & Guidelines for Quantification

ISO 27917:2017, (2017), Carbon dioxide Capture, Transportation and Geological Storage — Cross Cutting

ISO 9000:2005, (2005), Quality management systems — Fundamentals and vocabulary

ISO Guide 84:2020, (2020), Guidelines for Addressing Climate Change in Standards

ISO/IEC 14776:2010, (2010), Information technology — Small Computer System Interface (SCSI) — Part 121: Passive Interconnect Performance (PIP)

ISO/IEC 17000: 2004, (2004), ISO Standard- Conformity Assessement

ISOPA, (2012), Eco-profiles and Environmental Product Declarations of the European Plastics Manufacturer -Toluene Diisocyanate (TDI) Methylenediphenyl Diisocyanate (MDI), ISOPA

Jeswani, H., Krüger, C.; Kicherer, A., Anthony, F., Azapagic, A.; (2019), A Methodology for Integrating the Biomass balance approach into Lifecycle Assessment with an application in the Chemicals Sector; https://doi. org/10.1016/j.scitotenv.2019.06.088

Müller L.J., Kätelhön, A.; Bachmann, M., Zimmermann, A., Sternberg, A., Bardow, A.; (2020), A Guideline for Life Cycle Assessment of Carbon Capture and Utilization; https://doi.org/10.3389/fenrg.2020.00015

Net-Zero to Net-Negative, (2021), Net-Zero to Net-Negative: A Guide for Leaders on Carbon Removal.

PEF - Guide, (2012), Product Environmental Footprint Guide.

Plastics Europe - Steam Cracker Allocation, (2018), Plastics Europe recommendation on Steam Cracker allocation

Pathfinder Framework (PACT powered by WBCSD), (2022)

WBCSD - SOS 1.5, (2020), SOS 1.5- The Road to a Resilient, Zero Carbon Future

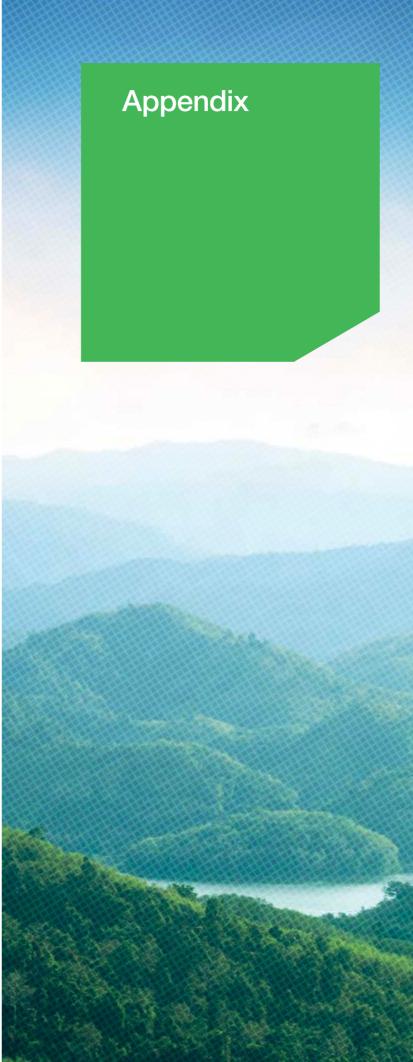
WBCSD, (2013), Guidance for Accounting and Reporting Corporate GHG Emissions in the Chemical Sector Value Chain, WBCSD Chemicals

WBCSD, (2014), Lifecycle Metrics for Chemical Products - A guideline by the chemical sector to assess and report on the environmental footprint of products, based on life cycle assessment

WBCSD, (2021), Pathfinder- Framework- Guidelines for the Accounting and Exchanging of Product Life Cycle Emissions

WBCSD, (2021), Reporting Matters

Winnipeg CO<sub>2</sub> Emission Factors, (), Emission factors in Kg CO<sub>2</sub>- Equivalent per Unit- Winnipeg https://www.winnipeg.ca/finance/findata/matmgt/documents/2012/682-2012/682-2012\_Appendix\_H-WSTP\_South\_End\_Plant\_Process\_Selection\_Report/Appendix%207.pdf; (Accessed 18 August 2022)



Proposals for calculating proxies in the case of no primary or secondary data are available

### **Example: Landfill**

The carbon content of the waste material shall be converted fully to CO<sub>2</sub>e when waste is disposed of on surface landfills.

There shall be no GHG emissions allocation for waste that is disposed of in underground landfills or similar (e.g. deep well injection).

- Waste to underground landfill: no GHG emissions to be allocated.
- Waste to surface landfill: 100% conversion to CO<sub>2</sub>e based on carbon content.

[BASF SE [2021]]

### **Example: Wastewater treatment**

Emissions from treatment of wastewater that is generated during the production of a product A be allocated to the PCF of the product A.

The GHG emissions calculation from wastewater treatment shall include the emissions coming from the biological degradation as well as the emissions from the operation of the wastewater treatment plant and the disposal of the sludge (incineration etc.). The carbon content of the waste material shall be converted fully to CO<sub>2</sub>e. As a basis for this calculation, the Total Organic Carbon (TOC) load of the process can be used if available.

If the Total Organic Carbon (TOC) load of your processes is known:

- 100% conversion to CO<sub>2</sub>e based on carbon content
- Utilities for treatment of wastewater and sludge incineration included using an emission factor of the treatment plant, e.g. 1 kg CO<sub>2</sub>e from treatment of 100 kg waste water.

[BASF SE [2021]]

e.g. A product generates 100 kg wastewater per kg of product. The amount of product therein is 0.1 kg.

0.001 kg  $\rm{CO_2e/kg}$  waste water from electricity

 $0.0005~{\rm kg~CO_2e/~kg}$  waste water from sludge incineration

PCF Product A = 0.001 kg CO $_2$ e/kg WWT electricity \* 100 kg + 0.0005 kg CO $_2$ e/kg WWT sludge incineration \* 100 kg + 0.7 kg CO $_2$ e/kg WWT TOC = 0.85 kg CO $_2$ e/kg

Further information can be found at:

Hernández-Chover, V.; Bellver-Domingo, A., Hernández-Sancho, F.; (2018), Efficiency of wastewater treatment facilities: The influence of scale economies, Journal of Environmental Management, Volume 228, 77-84, ISSN 0301-4797, https://doi.org/10.1016/j. jenvman.2018.09.014.



